

## Over one decade of invasion: the non-indigenous cladoceran *Evadne anonyx* G.O. Sars, 1897 in a low-salinity environment

Marilyn Kalaus\* and Henn Ojaveer

Estonian Marine Institute, University of Tartu, Lootsi 2a, 80012 Pärnu, Estonia

E-mail: [marilyn.kalaus@ut.ee](mailto:marilyn.kalaus@ut.ee) (MK), [henn.ojaveer@ut.ee](mailto:henn.ojaveer@ut.ee) (HO)

\*Corresponding author

Received: 4 March 2014 / Accepted: 12 August 2014 / Published online: 1 October 2014

Handling editor: Amy Fowler

### Abstract

Invasive species are often in focus in the non-indigenous species (NIS) research while low-abundance species receive relatively little attention. However, tracking NIS dynamics since the early stages of an invasion provides valuable information on the ecology of invasions. In the current paper, we investigated the invasion history and population dynamics of the small-bodied cladoceran *Evadne anonyx* G. O. Sars, 1897 in the Gulf of Riga (Baltic Sea) almost since its first detection in 2000. The species already was widespread the Gulf of Riga in 2001 and has been found in nearly every subsequent sample collected during summer months. However, the abundance of the species remained low, seldom 100 individuals  $m^{-3}$ . Both, salinity and water temperature affected the spatial distribution and population abundance of *E. anonyx*. The species was found to occur only sporadically at salinities below ca. 6. To obtain reliable presence/absence and density estimates on this small-bodied cladoceran, the entire zooplankton sample needed to be analysed.

**Key words:** distribution, invasion, Ponto-Caspian, population dynamics, sample analysis, Baltic Sea

### Introduction

Invasive species are a component of non-indigenous species (NIS) that can alter the structure and function of marine ecosystems and may result in appreciable socio-economic impacts to humans (Bax et al. 2003); consequently, they have become a special area of study. However, NIS that are currently rare might subsequently become invasive should environmental conditions become more suited for their expansion, but this might take several decades (e.g. Witte et al. 2010). Unfortunately, our current knowledge of invasive processes in many ecosystems is far from sufficient to predict invasive species expansions and their overall impacts with changing environmental conditions (Richardson 2011).

The study of a NIS population from an early invasion stage allows for a more complete understanding of the dynamics of such populations (O'Connor 2013; Alam et al. 2013). However, due to the scarcity of basic inventory studies in many marine waters (ICES 2012), marine invasions can go unnoticed, perhaps for several decades.

For example, the Ponto-Caspian hydromedusa *Blackfordia virginica* Mayer, 1910, remained unrecognized in South American waters for over fifty years (Freire et al. 2014). A second important factor is the lack of personnel with the ability to recognise and distinguish NIS from native organisms (e.g. Geller et al. 2010 and references therein).

To-date, 95 multicellular NIS have been recorded in the Baltic Sea since 1900 (Galil et al. 2014). Several of the most recent invasions (post 1990) originate from the Ponto-Caspian region, with transport on commercial vessels the most vector for introduction (AquaNIS 2013). The small cladoceran *Evadne anonyx* G.O. Sars, 1897, is one of these NIS and is native to the Caspian Sea (Rivier 1966) and estuarine areas of the Black and Azov Seas (Mordukhai-Boltovskoi and Rivier 1987; Rivier 1998). In the Baltic Sea, *E. anonyx* was first recorded in the Gulf of Finland in 1999 and a year later found in the Gulf of Riga (GoR) (Rodionova and Panov 2006; Pöllupüü et al. 2008). In 2006, the species was recorded in the Gulf of Gdansk (Bielecka et al. 2014), then two years

**Table 1.** Sampling details in the Gulf of Riga (GoR) and the outer station of Pärnu Bay (PB). Sampling depth in PB was 10m.

Year	Survey period		Station depth range (m)		Number of samples	
	GoR	PB	GoR	GoR	PB	
2001	19.07–24.07	-	26–54	10	-	
2002	25.07–31.07	-	20–25	10	-	
2006	26.07–31.07	08.06–02.10	20–45	13	34	
2007	24.07–30.07	03.05–11.10	16–40	9	20	
2008	23.07–29.07	09.05–17.09	20–35	16	16	
2009	22.07–28.07	15.05–10.08	10–30	13	13	
2010	21.07–27.07	25.05–31.08	10–40	20	14	
2011	26.07–01.08	10.05–22.09	22–45	19	25	
2012	25.07–31.07	09.05–09.10	9–46	22	21	

later in several parts of the open Baltic Sea (from the Arkona basin to the Landsort Deep and Askö) and has even extended its range into the Bothnian Sea (ICES 2009).

From the study of the long-term dynamics of eleven NIS in the NE Baltic Sea, it was clear that temperature significantly affected abundance/biomass of most of these species (Ojaveer et al. 2011). *E. anonyx* typically appears in the mesozooplankton community in May, reaches peak abundance during July, and disappears in October (Rodionova and Panov 2006; Pöllupüü et al. 2008). The species rarely exceeds 100 individuals  $m^{-3}$  (Pöllupüü et al. 2008) but is consistently present when water temperatures exceed 12–15°C and salinity is >5, with a maximum abundance at 17–19°C (Rodionova and Panov 2006; Pöllupüü et al. 2008). Within its native environment (Caspian Sea), *E. anonyx* abundance peaks at temperatures of 16–20°C and salinity of 12–13 (Mordukhai-Boltovskoi and Rivier 1987; Rivier 1998). Thus, the salinity and temperature conditions are important controlling the seasonal population dynamics of *E. anonyx*.

The aim of this study was to investigate: i) the seasonal and long-term abundance of *E. anonyx* in one of the distinct sub-basins of the Baltic Sea (GoR) with sampling having commenced from about the date of its first detection; and ii) how its distribution and abundance might be influenced by variation salinity and water temperature.

## Materials and methods

### Study area

The GoR has an area of 16 330 km<sup>2</sup> and a volume 424 km<sup>3</sup>. It is a relatively shallow (mean depth 26 m and maximum depth >60 m) and semi-enclosed sub-basin in the NE part of the

Baltic Sea. The basin receives freshwater from a large drainage area (134,000 km<sup>2</sup>), which mostly enters at the southern part of the basin. Salinity varies from 0.5 to 7 and there is no permanent halocline. Oxygen concentrations above 5 mL L<sup>-1</sup> in most areas. In summer, the surface temperature may attain 20°C (Kotta et al. 2008).

Pärnu Bay (PB) has an area of 700 km<sup>2</sup> and a volume 2 km<sup>3</sup>. This is a shallow area with an average depth of 5 m in the NE part of the GoR. PB is ice-covered during most winters. During the ice-free season, the water column is well mixed. During July and August, surface temperatures may rise to 24°C, and salinity varies from almost fresh water in the Pärnu River mouth to 7 in the open bay. Hydrological conditions are determined by meteorological processes, river discharge (Pärnu River, freshwater inflow 2 km<sup>3</sup> annually), and water exchange with the open part of the GoR (Kotta et al. 2008).

### Sampling and sample analysis

Zooplankton samples were collected during day-time with a Juday net (mouth area 0.1 m<sup>2</sup>, 90 µm bar-mesh) hauled vertically from the bottom to the surface. Samples were preserved in a 4% formalin and seawater solution. In the open GoR, zooplankton samples were obtained during hydro-acoustic surveys for pelagic commercial fish performed mainly in July during 2001–2012 (see Table I). There was no sampling during 2003 to 2005. Sampling in PB was at two stations, located in the inner and outer part (depths 5 and 10m, respectively), during May to October of 2006–2012 (Figure 1). Results of zooplankton sampling prior to 2006 are reported by Pöllupüü et al. (2008) and these data were included in the present study. Samples from station 2 in PB were obtained bimonthly during June–August, whereas

sampling in May and September–October took place monthly. At PB station 1, samples were collected monthly. In all, 132 samples were collected from the open part of the GoR and 143 samples from PB (Table I and Supplementary material Tables S1–S2). To obtain the abundance of *E. anonyx*, each sample was examined in its entirety.

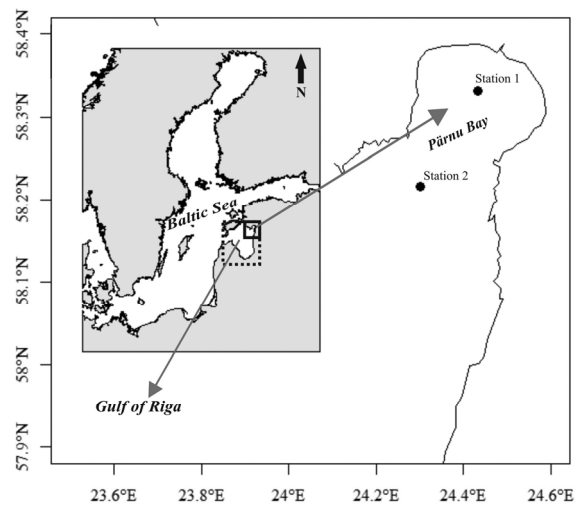
In PB, temperature and salinity were measured throughout the water column during zooplankton surveys using CTD (model SAIV SD204; SAIV A/S, Environmental Sensors & Systems, Laksevag, Norway). For the GoR, we used the mean sea-surface temperature (SST) and sea-surface salinity (SSS) values for each survey period, determined from the daily modelled data obtained from Omstedt (2011, updated).

The native *Evadne nordmanni* Lovén, 1836, and the alien *E. anonyx* are routinely examined in PB as part of the Estonian marine monitoring program using HELCOM COMBINE recommendations. This methodology is based on sub-sample analysis of zooplankton samples (HELCOM 2013). To obtain reliable estimates of infrequent small-bodied NIS, such as *E. anonyx*, we performed a comparative analysis and compared the efficacy of the total sampling method and the sub-sampling method. We scored all *E. anonyx* and *E. nordmanni* separately in sub-samples and then calculated the percentage of *E. anonyx* to those of the genus *Evadne* in each sub-sample. We then applied this ratio to the genus *Evadne* to obtain an abundance estimate obtained for the total sample and then compared these with the total-sample count and both expressed as the number of individuals  $m^{-3}$ .

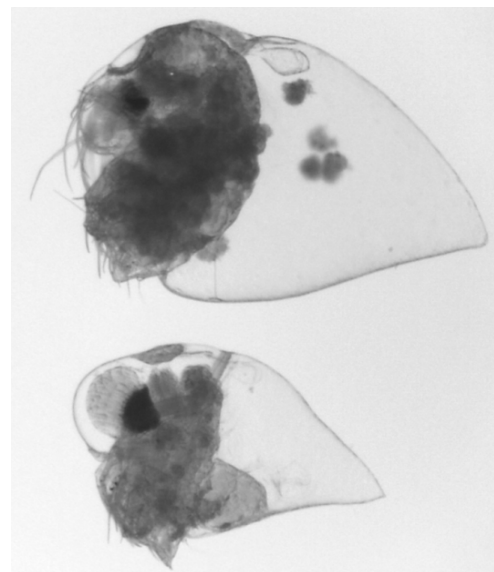
Morphologically, *E. anonyx* and *E. nordmanni* are very similar (Figure 2). The distinction was made using two characteristics outlined by Rivier (1998): (i) formula of the setae – number of setae in the exopodites of thoracic limbs I–IV – for *E. anonyx* 2.2.2.1 and for *E. nordmanni* 2.2.1.1; and (ii) shape of cauda – for *E. anonyx* the cauda is formed of rounded tubercles and the cauda of *E. nordmanni* has two conical pointed outgrowths.

#### Statistical analysis

The analyses were undertaken using the open-source language R version 3.0.2 (R Core Team 2013). Log-transformed abundances of *E. anonyx* were modelled as a response to water temperature and salinity. Temperature and abundance was expected to show a seasonal and interannual variability, which we separated by fitting the generalized additive model (GAM) smoother (R

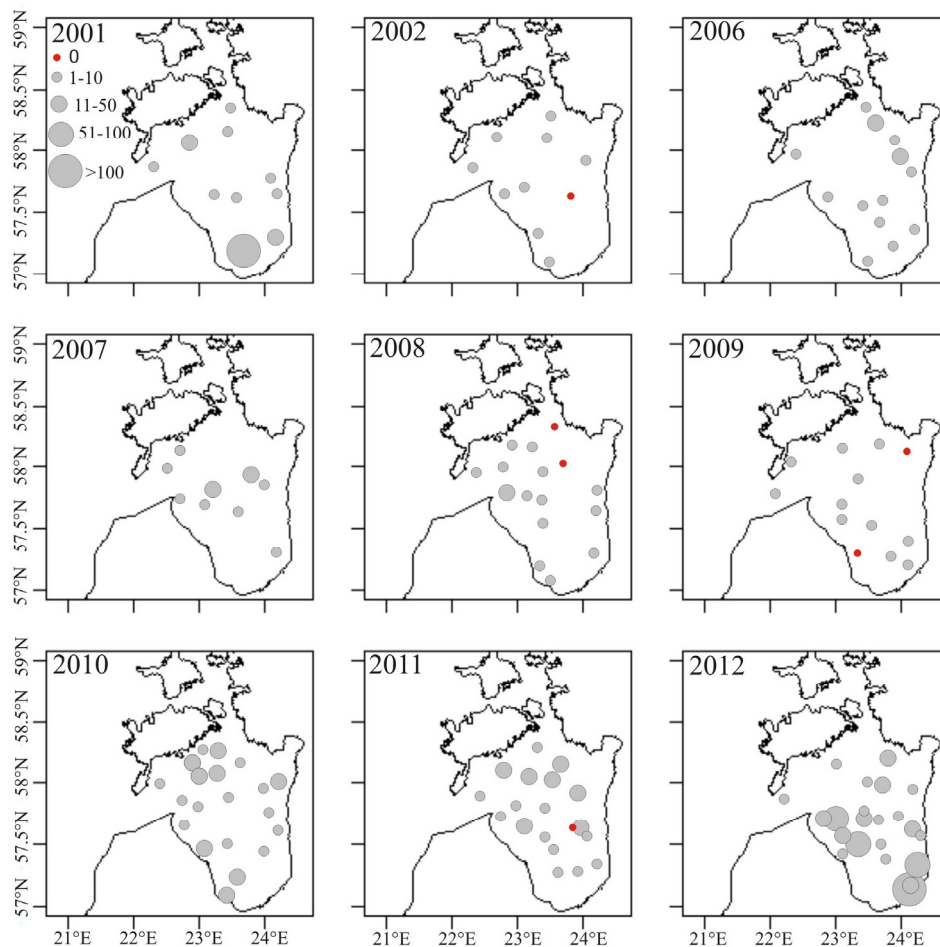


**Figure 1.** Location of the study area: Gulf of Riga, and inner (1) and outer (2) station in Pärnu Bay (NE Gulf of Riga).



**Figure 2.** *Evadne anonyx* (upper) and *Evadne nordmanni* (lower). Photo by M. Kalas.

library *mgcv*; Wood 2006) through abundance and the temperature, using the week of year as a predictor variable. A linear model was fitted using either the GAM predicted seasonal values of abundance and the temperature, or residuals from the seasonal GAMs. Correlation in the first (predicted seasonal values of abundance and temperature) indicated the effect of seasonality. The effect of interannual temperature variability on abundance is present should a correlation



**Figure 3.** Spatial distribution and abundance (individuals  $\text{m}^{-3}$ ) of the non-indigenous *Evadne anonyx* in the Gulf of Riga in summer 2001–2012.

between the residuals of seasonal abundance and the temperature exist. Due to sample/data availability, the seasonal relations between the abundance of *E. anonyx* and abiotic environment were investigated for the PB station 2 for June–August period only. Comparisons between the two different methods for analysing samples involved the Student's *t*-test. Non-zero log-transformed abundances were included, resulting in normal distribution of two samples compared.

## Results

### *Spatio-temporal distribution of Evadne anonyx*

*E. anonyx* was already widely distributed in the GoR during 2001 – it was present in 126 of 132 samples. The highest abundances occurred in the

southern and western parts of the basin (Figure 3; Supplementary material Table S1). During the period 2002 to 2009, mean abundance was somewhat lower than in 2001 (Figure 4). The mean abundance ranged between 3 and 28 ind.  $\text{m}^{-3}$  with the highest density (173 ind.  $\text{m}^{-3}$ ) observed in 2012. Only parthenogenetic females were captured in the GoR samples in summer. When present, *E. anonyx* comprised 0.4–100% of the total abundance of the genus *Evadne*, with a mean 42.7% and standard deviation of 36.6%.

Few *E. anonyx* were obtained from PB samples before 2006 (M. Põllupüü, University of Tartu, Tallinn, Estonia, unpubl. data). Within the inner part of this bay (station 1), a similar situation continued during our study period, we only detected the species in two years (2006 and 2011). In 2006, *E. anonyx* occurred mainly during August with

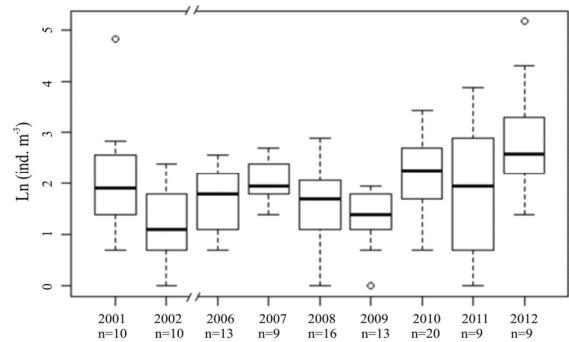
abundances of 2 to 50 ind. m<sup>-3</sup>. In July and August 2011, we observed densities of 2 to 44 ind. m<sup>-3</sup>. At PB station 2, the abundance of *E. anonyx* was relatively low, although higher than at station 1. *E. anonyx* was found during May to October in all years investigated at densities of 1 to 22 ind. m<sup>-3</sup> and was more abundant during 2006–2008. Population numbers declined to a low level in 2009–2011 (Figure 5; Table S2). The highest abundance in a single sample (153 ind. m<sup>-3</sup>) was recorded in August, 2006. Unlike the summer samples in the GoR, both partenogenetic and gamogenetic individuals occurred (see also Põllupüü et al. 2008).

*Relation of E. anonyx abundance to environmental conditions*

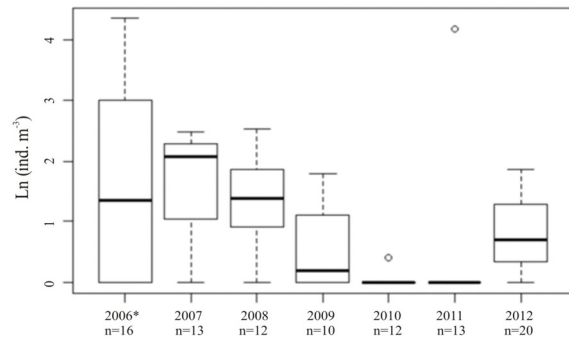
There was a highly significant positive correlation between the abundance of *E. anonyx* and SSS for the GoR ( $R^2=0.223$ ,  $p<0.001$ , GAM; SSS range 6.1–6.7). The relationship with SST was also statistically significant ( $R^2=0.11$ ,  $p=0.02$ , GAM; SST range 16.1–20.8 °C). In the outer part of PB at station 2, we found a positive relationship between the abundance of *E. anonyx* and water temperature ( $R^2=0.241$ ,  $p<0.001$ , GAM; temperature range 4.0–25.6°C). There was no correlation between the residuals of the seasonal GAM curve of abundance and temperature, i.e. the positive linear correlation between the *E. anonyx* abundance and temperature was due to the similar seasonal course of both variables. The response on species abundance was slightly lagged and followed seasonal changes in temperature, both with increasing temperature in spring/summer followed by decreasing temperature in late summer to early autumn (Figure 6). We found no relation between the abundance of *E. anonyx* and salinity in the outer PB ( $R^2=0.095$ ,  $p=0.219$ ; salinity range 3.1–5.6).

*Methods comparison*

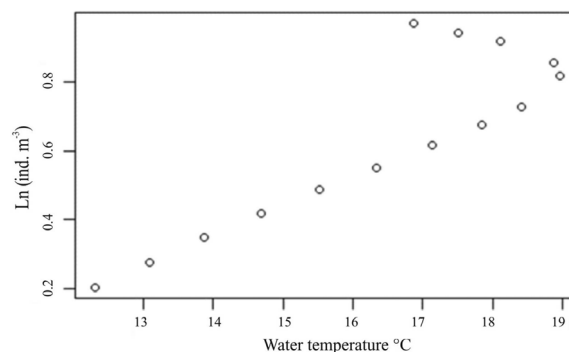
The two methods to estimate *E. anonyx* abundance were not equivalent. Subsampling only detected *E. anonyx* in 55 (46 %) of zooplankton samples where the species was actually present as determined by total count. As well, subsampling overestimated the *E. anonyx* abundance in samples when detected ( $t=-4.009$ ,  $df=35$ ,  $p<0.01$ ,  $n=36$ , t-test; Figure 7). As a result the sub-sample method was judged unreliable for estimating presence and/or densities of *E. anonyx*.



**Figure 4.** Abundance (individuals m<sup>-3</sup>) of *Evadne anonyx* in the Gulf of Riga in summer during 2001–2012. On each box, the central mark is the median, boxes indicate first and third quartiles, whiskers are 95% confidence bounds, and dots represent outliers.



**Figure 5.** Abundance (individuals m<sup>-3</sup>) of *Evadne anonyx* in the outer station of Pärnu Bay in years 2006–2012 (\* data from M. Põllupüü, University of Tartu, Tallinn, Estonia, unpublished data.). On each box, the central mark is the median, boxes indicate first and third quartiles, whiskers are 95% confidence bounds, and dots represent outliers.



**Figure 6.** GAM predicted values of *Evadne anonyx* abundance vs. GAM predicted seasonal curve of water temperature in the outer Pärnu Bay. Both species abundance and water temperature had weekly resolution.

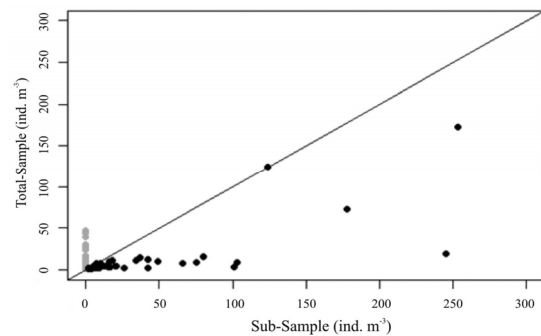
## Discussion

The Baltic NIS has been colonized by several pelagic invertebrates during the past few decades. In addition to *E. anonyx*, *Cercopagis pengoi* Ostroumov, 1891, *Mnemiopsis leidyi* A. Agassiz, 1865, and *Cornigerius maeoticus* Pengo, 1879, have arrived (AquaNIS 2013). The current study provides evidence of a seasonal and spatio-temporal dynamics of *E. anonyx* since the early invasion stage. *E. anonyx* has formed self-sustaining population across the Gulf of Riga since 2001, or perhaps before, and can now be considered to be a common component of the mesozooplankton community. However, densities of *E. anonyx* have remained relatively low throughout the study period.

Pöllupüü et al. (2008) stated that *E. anonyx* was not considered to be a high risk species for low-salinity environments in the Baltic Sea because it was suggested that the species is unable to survive at salinities below 9 (Panov et al. 1999; Rivier 1998). Based on the current study, the species can be considered to be a permanent component of the mesozooplankton community above salinities ~6. However, *E. anonyx* clearly suffers under substantial salinity stress across the whole GoR. We also suggest that the somewhat sporadic occurrence of *E. anonyx* in the inner part of PB, where salinity is generally below 6, may be due wind-induced water currents that transport *E. anonyx* from the open GoR. Salinity stress is a feature challenging other NIS in the Baltic Sea, as in case of the comb jelly *Mnemiopsis leidyi* (Jaspers et al. 2011) and Chinese mitten crab *Eriocheir sinensis* H. Milne-Edwards, 1853 (Otto and Brandis 2011).

A recent study in the Gulf of Gdansk in the southern Baltic Sea failed to show a relationship between *E. anonyx* abundance (variability 0.11–6 ind. m<sup>-3</sup>) and salinity (range 4.6–7.5). When compared with our study, the abundance of *E. anonyx* in the Gulf of Gdansk was substantially lower although the salinity range was greater (Bielecka et al. 2014). This might be explained due to the low density of the *E. anonyx* and insufficient spatio-temporal coverage. However, the species may have only recently arrived to this region (Bielecka et al. 2014). In terms of water temperature, the highest densities of *E. anonyx* in the Gulf of Gadansk (Pöllupüü et al. 2008; Bielecka et al. 2014) occurred at the same 19°C temperature as observed in the current study.

Cladoceran populations have high seasonal variability, partly due to the alternation of partheno-



**Figure 7.** Comparison of the abundance estimates of the non-indigenous cladoceran *Evadne anonyx* based on Sub-sampling and total sample counts. Grey dots indicate samples, where sub-sampling failed to detect *E. anonyx*

genetic and gamogenetic reproduction modes, and are characterised by a rapid increase in densities when environmental conditions became favourable (Egloff et al. 1997; Rivier 1998). High population abundances are generally sustained only for a relatively short time. Earlier investigations have confirmed that parthenogenetic females dominated in the *E. anonyx* community in the GoR for most of the season (Pöllupüü et al. 2008). The decrease in abundance in autumn coincided with the increasing abundance of males and gamogenetic females (Pöllupüü et al. 2008). It has been suggested that shift to a gamogenetic reproduction mode is triggered by unfavourable environmental conditions (Egloff et al. 1997; Rivier 1998). As we have found only parthenogenetic *E. anonyx* individuals in the GoR samples in summer, the population was therefore still in the developing phase in all years investigated and likely had not reached the annual abundance peak.

In contrast to *E. anonyx*, the non-indigenous predatory cladoceran *Cercopagis pengoi* may attain greater abundances in the GoR, principally within sheltered coastal areas. The seasonal timing of peak abundance *E. anonyx* broadly matches that of *C. pengoi* (Pöllupüü et al. 2008; Ojaveer et al. 2004). As small-bodied, low-motility, species are preferred prey for *C. pengoi* (Simm and Ojaveer 2006), *E. anonyx* may be a valuable prey item for *C. pengoi* during the warmest summer months - a time when the preferred prey *Eubosmina* spp. is less abundant (Ojaveer et al. 2004). *E. anonyx* might therefore facilitate other pelagic alien species in the area as a prey item. *Evadne* spp. are seldom consumed by pelagic fishes in the GoR during summer (Lankov et al. 2010), possibly due to their low densities; however, they have been found in the Baltic herring (*Clupea harengus membras* Linnaeus, 1761) stomachs (M. Kalaus,



unpubl. data). However, *C. pengoi* may play an important role in herring and three-spine stickleback (*Gasterosteus aculeatus* Linnaeus, 1758) diets in the GoR (Lankov et al. 2010). We suggest the ecological role of *E. anonyx* in the GoR, due to its low abundance, is most likely limited to enriching the species diversity of pelagic communities during the warm season.

As ecosystem conditions may change, the status of a NIS population might change over time. For example, a complex of changes in the environment resulted in a sudden outburst of the comb jelly *Mnemiopsis leidyi* in the Black Sea and led to significant decline in abundance of pelagic commercial fishes (e.g. Oguz et al. 2008). Therefore, systematic monitoring of NIS population dynamics, and investigating their ecology over time, is essential to track and explain potential changes in ecosystem structure and functioning. The current study provides a baseline of the seasonality and abundance levels for *E. anonyx* in the GoR under conditions characterised by a relatively warm climate and low salinity. While the densities of *E. anonyx* remained, in some instances, several orders of magnitude lower than those of many native zooplankton species (e.g., Ikauniece 2001; Lankov et al. 2010) this situation might change if environmental conditions change.

Successful reproduction and establishment of a self-sustaining NIS population under the lower levels of salinity might indicate an osmoregulatory adaptation to local conditions. Physiological plasticity has indeed shown to be an important characteristic of several NIS (e.g. Blumenshine et al. 2011 and references therein). To our knowledge, there is no indication that this has happened for *E. anonyx*. However, unexpected low-salinity preference was recently observed for an invasive gammarid *Gammarus tigrinus* Sexton, 1939 in the Baltic Sea, potentially due to high genetic diversity of invading individuals (Kotta et al. 2013). Thus, further laboratory experiments and investigations in physiology and genetics of *E. anonyx* is needed to shed some light on the causes and potential future /consequences of this colonization.

We also found that the entire mesozooplankton sample should be examined when scoring the numbers of infrequent small-bodied specimens as sub-sampling may either not capture or poorly estimated numbers of individuals. This is a practical finding that relates to meeting international requirements like EU MSFD Descriptor 2 (European Commission 2010) and when performing basin-scale impact assessments of NIS (Olenin et al. 2007).

## Acknowledgements

We thank three anonymous referees for their helpful comments on an earlier version of the manuscript and D. Minchin for comments on the text. This work was partially financed by Estonian Ministry of Education and Research (grant SF0180005s10) and the Estonian Science Foundation (grant 8747). The research leading to the results has also received funding from the European Community's Seventh Framework Program (FP7/2007–2013) under Grant Agreement No. 266445 for the Vectors Change in Oceans and Seas marine Life, Impact on Economics Sectors (VECTORS). This work was also co-financed by the project: "The status of marine biodiversity and its potential futures in the Estonian coastal sea 3.2.0801.11-0029" of Environmental protection and technology programme of the European Regional Fund. The paper contributes to the project "Biodiversity changes - causes, consequences and management implications" (BIO-C3) of the joint Baltic Sea research and development programme (Art 185) BONUS, funded jointly from the European Union's Seventh Programme for research, technological development and demonstration and from Estonian Research Council.

## References

- Alam Rashidul AKM, Hagino T, Fukaya K, Okuda T, Nakaoka M, Noda T (2013) Early phase of the invasion of *Balanus glandula* along the coast of Eastern Hokkaido: changes in abundance, distribution, and recruitment. *Biological Invasions* 16: 1699–1708, <http://dx.doi.org/10.1007/s10530-013-0619-4>
- AquaNIS (2013) Information System on Aquatic Non-indigenous and Cryptogenic Species <http://www.corpi.ku.lt/databases/index.php/aquanis> (Accessed 11 February 2014)
- Bax N, Williamson A, Aguero M, Gonzalez E, Geeves W (2003) Marine invasive alien species: a threat to global biodiversity. *Marine Policy* 27: 313–323, [http://dx.doi.org/10.1016/S0308-597X\(03\)00041-1](http://dx.doi.org/10.1016/S0308-597X(03)00041-1)
- Bielecka L, Mudrak-Cegiołka S, Kalarus M (2014) *Evadne anonyx* G. O. Sars, 1897 – the first record of this Ponto-Caspian cladoceran in the Gulf of Gdańsk (Baltic Sea). *Oceanologia* 56: 141–150, <http://dx.doi.org/10.5697/oc.56-1.141>
- Blumenshine SC, Tsukimura B, Rice A, Rudnick DA (2011) Environmental factors influencing the dynamics of Chinese mitten crab zoea in the San Francisco Bay-Delta. *Aquatic Invasions* 7: 111–124, <http://dx.doi.org/10.3391/ai.2012.7.1.012>
- Egloff DA, Fofonoff PW, Onbé T (1997) Reproductive Biology of Marine Cladocerans. *Advances in Marine Biology* 31: 79–167, [http://dx.doi.org/10.1016/S0065-2881\(08\)60222-9](http://dx.doi.org/10.1016/S0065-2881(08)60222-9)
- European Commission (2010) Decision on Criteria and Methodological Standards on Good Environmental Status of Marine Waters. Decision 2010/477/EU. Official Journal of the European Union L232: 14–24
- Freire M, Genzano GN, Neumann-Leitão S, Pérez CD (2014) The non-indigenous medusa *Blackfordia virginica* (Hydrozoa, Leptothecata) in tropical Brazil: 50 years of unnoticed presence. *Biological Invasions* 16: 1–5, <http://dx.doi.org/10.1007/s10530-013-0496-x>
- Galil BS, Marchini A, Occhipinti-Ambrogi A, Mitchin D, Narščius A, Ojaveer H, Olenin S (2014) International arrivals: widespread bioinvasions in European Seas. *Ethology Ecology & Evolution* 26: 152–171, <http://dx.doi.org/10.1080/03949370.2014.897651>
- Geller JB, Darling JA, Carlton JT (2010) Genetic perspectives on marine biological invasions. *Annual Review of Marine Science* 2: 367–394, <http://dx.doi.org/10.1146/annurev.marine.010908.163745>

- HELCOM COMBINE (2013) <http://helcom.fi/Documents/Action%20areas/Monitoring%20and%20assessment/Manuals%20and%20Guidelines/Manual%20for%20Marine%20Monitoring%20in%20the%20COMBINE%20Programme%20of%20HELCOM.pdf> (Accessed 28 October 2013)
- ICES (2012) Report of the ICES Working Group on Introduction and Transfers of Marine Organisms (WGITMO), 14–16 March 2012, Lisbon, Portugal. ICES CM 2012/ACOM:31, 301 pp
- ICES (2009) Report of the Working Group on Introduction and Transfers of Marine Organisms (WGITMO), 11–13 March 2009, Washington D.C., USA. ICES CM 2009/ACOM:25, 220 pp
- Ikaunieca A (2001) Long-term abundance dynamics of coastal zooplankton in the Gulf of Riga. *Environment International* 26: 175–181, [http://dx.doi.org/10.1016/S0160-4120\(00\)00094-5](http://dx.doi.org/10.1016/S0160-4120(00)00094-5)
- Jaspers C, Møller LF, Kiørboe T (2011) Salinity Gradient of the Baltic Sea Limits the Reproduction and Population Expansion of the Newly Invaded Comb Jelly *Mnemiopsis leidyi*. *PLoS ONE* 6: e24065, <http://dx.doi.org/10.1371/journal.pone.0024065>
- Kotta J, Pärnoja M, Katajisto T, Lehtiniemi M, Malavin SA, Reialu G, Panov VE (2013) Is a rapid expansion of the invasive amphipod *Gammarus tigrinus* Sexton, 1939 associated with its niche selection: a case study in the Gulf of Finland, the Baltic Sea. *Aquatic Invasions* 8: 319–332, <http://dx.doi.org/10.3391/ai.2013.8.3.08>
- Kotta J, Lauringson V, Martin G, Simm M, Kotta I, Herkül K, Ojaveer H (2008) Gulf of Riga and Pärnu Bay. In: Schiewer U (ed), Ecology of Baltic Coastal waters, Ecological Studies, vol. 197, Springer-Verlag, Berlin Heidelberg, pp 217–243, [http://dx.doi.org/10.1007/978-3-540-73524-3\\_10](http://dx.doi.org/10.1007/978-3-540-73524-3_10)
- Lankov A, Ojaveer H, Simm M, Pöllupüü M (2010) Feeding ecology of pelagic fish species in the Gulf of Riga (Baltic Sea): the importance of changes in the zooplankton community. *Journal of Fish Biology* 77: 2268–2284, <http://dx.doi.org/10.1111/j.1095-8649.2010.02805.x>
- Mordukhai-Boltovskoi FD, Rivier IK (1987) Predatory Cladocerans Podonidae, Polyphemidae, Cercopagidae and Leptodoridae of the World Fauna. Nauka Publishing, Moscow, 183 pp
- O'Connor NJ (2013) Invasion dynamics on a temperate rocky shore: from early invasion to establishment of a marine invader. *Biological Invasions* 16: 73–87, <http://dx.doi.org/10.1007/s10530-013-0504-1>
- Oguz T, Fach B, Salioglu B (2008) Invasion dynamics of the alien ctenophore *Mnemiopsis leidyi* and its impact on anchovy collapse in the Black Sea. *Journal of Plankton Research* 30: 1385–1397, <http://dx.doi.org/10.1093/plankt/fbn094>
- Ojaveer H, Kotta J, Pöllumäe A, Pöllupüü M, Jaanus A, Vetemaa M (2011) Alien species in a brackish water temperate ecosystem: Annual-scale dynamics in response to environmental variability. *Environmental Research* 111: 933–942, <http://dx.doi.org/10.1016/j.envres.2011.03.002>
- Ojaveer H, Simm M, Lankov A (2004) Population dynamics and ecological impact of the non-indigenous *Cercopagis pengoi* in the Gulf of Riga (Baltic Sea). *Hydrobiologia* 522: 261–269, <http://dx.doi.org/10.1023/B:HYDR.0000029927.91756.41>
- Olenin S, Minchin D, Daunys D (2007) Assessment of biopollution in aquatic ecosystems. *Marine Pollution Bulletin* 55: 379–394, <http://dx.doi.org/10.1016/j.marpolbul.2007.01.010>
- Omstedt A (2011) Guide to process based modelling of lakes and coastal seas. Springer-Praxis books in Geophysical Sciences, Springer-Verlag Berlin Heidelberg, 281 pp, <http://dx.doi.org/10.1007/978-3-642-17728-6>
- Otto T, Brandis D (2011) First evidence of *Eriocheir sinensis* reproduction from Schleswig-Holstein, Northern Germany, western Baltic Sea. *Aquatic Invasions* 6: 65–69, <http://dx.doi.org/10.3391/ai.2011.6.S1.015>
- Panov VE, Krylov PI, Telesh IV (1999) The St. Petersburg harbour profile. In: Gollasch S, Leppäkoski E (eds), Initial Risk Assessment of Alien Species in Nordic Coastal Waters. Nord 8. Nordic Council of Ministers, Copenhagen, 245 pp
- Pöllupüü M, Simm M, Pöllumäe A, Ojaveer H (2008) Successful establishment of the Ponto-Caspian alien cladoceran *Evadne anonyx* G.O. Sars 1897 in low-salinity environment in the Baltic Sea. *Journal of Plankton Research* 30: 777–782, <http://dx.doi.org/10.1093/plankt/fbn036>
- R Core Team (2013) R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL: <http://www.r-project.org>
- Richardson DM (2011) Invasion science: the roads travelled and the roads ahead. In: Richardson DM (ed), Fifty Years of Invasion Ecology: The Legacy of Charles Elton, 1st edition. Blackwell Publishing Ltd, pp 397–407
- Rivier IK (1998) The Predatory Cladocera (Onychopoda: Podonidae, Polyphemidae, Cercopagidae) and Leptodoridae of the World. Backhuys Publishers Leiden, The Netherlands, pp 132–137
- Rivier IK (1966) K taksonomij *Evadne anonyx* Sars. Trudy Instituta Biologii Vnutrennikh Vod 12, 151–158 (in Russian)
- Rodionova NV, Panov VE (2006) Establishment of the Ponto-Caspian predatory cladoceran *Evadne anonyx* in the eastern Gulf of Finland, Baltic Sea. *Aquatic Invasions* 1: 7–12, <http://dx.doi.org/10.3391/ai.2006.1.1.3>
- Simm M, Ojaveer H (2006) Taxonomic status and reproduction dynamics of the non-indigenous *Cercopagis* in the Gulf of Riga (Baltic Sea). *Hydrobiologia* 554: 147–154, <http://dx.doi.org/10.1007/s10750-005-1015-6>
- Witte S, Buschbaum C, van Beusekom JEE, Reise K (2010) Does climatic warming explain why an introduced barnacle finally takes over after a lag of more than 50 years? *Biological Invasions* 12:3579–3589, <http://dx.doi.org/10.1007/s10530-010-9752-5>
- Wood SN (2006) Generalized Additive Models. An Introduction with R.; Carlin BP, Chatfield C, Tanner M, Zidek J, editors: Chapman and Hall/CRC, 392 pp

The following supplementary material is available for this article:

**Table S1.** Abundance of *Evadne anonyx* in the Gulf of Riga (by sampling dates and stations) in summer, 2001–2012.

**Table S2.** Abundance of *Evadne anonyx* in the outer part of Pärnu Bay (station 2, see Figure 1) by sampling dates in 2006–2012.

This material is available as part of online article from: [http://www.aquaticinvasions.net/2014/Supplements/A1\\_2014\\_Kalaus\\_Ojaveer\\_Supplement.xls](http://www.aquaticinvasions.net/2014/Supplements/A1_2014_Kalaus_Ojaveer_Supplement.xls)