

# Changes in microbial communities in coastal sediments along natural CO<sub>2</sub> gradients at a volcanic vent in Papua New Guinea

Felix F. Raulf,<sup>1</sup> Katharina Fabricius,<sup>2</sup> Sven Uthicke,<sup>2</sup> Dirk de Beer,<sup>3</sup> Raeid M. M. Abed<sup>4</sup> and Alban Ramette<sup>1\*</sup>

<sup>1</sup>HGF-MPG Joint Research Group on Deep Sea Ecology and Technology and <sup>3</sup>Microsensor Group, Max Planck Institute for Marine Microbiology, Bremen, Germany.

<sup>2</sup>Water Quality and Ecosystem Health, Australian Institute of Marine Science, Townsville, Australia.

<sup>4</sup>Department of Biology, Sultan Qaboos University, Muscat, Oman.

## Summary

**Natural CO<sub>2</sub> venting systems can mimic conditions that resemble intermediate to high pCO<sub>2</sub> levels as predicted for our future oceans. They represent ideal sites to investigate potential long-term effects of ocean acidification on marine life. To test whether microbes are affected by prolonged exposure to pCO<sub>2</sub> levels, we examined the composition and diversity of microbial communities in oxic sandy sediments along a natural CO<sub>2</sub> gradient. Increasing pCO<sub>2</sub> was accompanied by higher bacterial richness and by a strong increase in rare members in both bacterial and archaeal communities. Microbial communities from sites with CO<sub>2</sub> concentrations close to today's conditions had different structures than those of sites with elevated CO<sub>2</sub> levels. We also observed increasing sequence abundance of several organic matter degrading types of *Flavobacteriaceae* and *Rhodobacteraceae*, which paralleled concurrent shifts in benthic cover and enhanced primary productivity. With increasing pCO<sub>2</sub>, sequences related to bacterial nitrifying organisms such as *Nitrosococcus* and *Nitrospirales* decreased, and sequences affiliated to the archaeal ammonia-oxidizing *Thaumarchaeota Nitrosopumilus maritimus* increased. Our study sug-**

**gests that microbial community structure and diversity, and likely key ecosystem functions, may be altered in coastal sediments by long-term CO<sub>2</sub> exposure to levels predicted for the end of the century.**

## Introduction

Two hundred years of anthropogenic activities, such as deforestation, agricultural land use and most prominently the use of fossil carbon as energy source, have led to an increase in atmospheric CO<sub>2</sub> concentrations, from long-term stable levels of around 280 μatm pCO<sub>2(atm)</sub> to presently > 400 μatm (Stocker *et al.*, 2013). The dissolution of atmospheric CO<sub>2</sub> in seawater significantly increases concentrations of dissolved inorganic carbon in the ocean and has reduced mean surface seawater pH by 0.1 units, a process commonly known as ocean acidification (OA). Realistic CO<sub>2</sub> emission scenarios of the Intergovernmental Panel for Climate Change (Stocker *et al.*, 2013) predict atmospheric CO<sub>2</sub> levels of 750 μatm and higher by the end of the century, which will result in a further pH drop of up to 0.4 units (Raven *et al.*, 2005; Andersson *et al.*, 2007). This change would be accompanied by a significant decrease in seawater carbonate ion concentrations and a lower saturation state for calcium carbonate minerals, directly affecting marine calcareous organisms such as coccolithophorids, foraminifera, scleractinian corals and coralline algae (Hall-Spencer *et al.*, 2008; Fabricius *et al.*, 2011; Uthicke *et al.*, 2013). Current OA has already measurable negative impacts on coral ecosystems and this trend will worsen in the future (Pandolfi *et al.*, 2011).

Benthic microbial communities, especially in the oxic layer of the surface of coastal marine sediments, may also be directly influenced by changes in seawater chemistry, potentially leading to different functional responses. Microbial communities play a critical role for ecosystems by supporting the main biogeochemical cycles, such as primary production and remineralization of organic material (Decho, 2000; Thompson *et al.*, 2004; Hewson *et al.*, 2007), by providing protection to larger organisms as biofilms (Mouchka *et al.*, 2010), and/or by influencing the settlement of faunal larvae on benthic substrata

Received 27 June, 2014; revised 12 November, 2014; accepted 23 November, 2014. \*For correspondence. E-mail alban.ramette@ispm.unibe.ch, Institute of Social and Preventive Medicine, University of Bern, Finkenhubelweg 11, 3012 Bern, Switzerland; Tel. +41 31 631 5975; Fax +41 31 631 3648.

(Webster *et al.*, 2004). Despite a growing understanding of the structure of shallow sediment microbiota and of their role in the dynamics of coastal sediment ecosystems (e.g. Böer *et al.*, 2009), responses of benthic microbial communities to environmental changes such as OA are still poorly understood.

Recent experimental studies on bacterioplankton and open water microbial assemblages indicated that increased OA might affect microbial processes such as nitrification rates (Beman *et al.*, 2011) or activities of certain extracellular enzymes (Piontek *et al.*, 2009) and could significantly alter bacterial community structure in the water column (Krause *et al.*, 2012). Yet, all existing experiments and studies related to OA effects on microbial communities have, so far, only investigated responses to exposure that lasted days to weeks, and only one study exists on long-term effects on community composition and structure in coastal sediments (Kitidis *et al.*, 2011). Community responses to increased  $p\text{CO}_2$  observed in short-term experiments obviously do not reflect long-term acclimatization, evolutionary adaptations, complex feedbacks or indirect effects (e.g. changes in local geochemistry, faunal and floral composition), which occur within a natural marine system subjected to decennia or centuries of exposure. To obtain a realistic description of future impacts of chronic OA, ecosystems that are naturally exposed to high  $p\text{CO}_2$  must therefore be studied.

Here, we examined the composition and diversity of bacterial and archaeal communities associated with benthic sediments along a natural  $p\text{CO}_2$  gradient, formed by volcanic  $\text{CO}_2$  venting. The local conditions provide long-term  $p\text{CO}_2$  gradients in the range of acidification projections for the next century, a high purity of the emitted volcanic gas as well as limited confounding effects caused by fluctuations in temperature, salinity or strong currents (Fabricius *et al.*, 2011). Although many coastal systems experience natural pH fluctuations due to seasonal or annual changes or to microbial activities such as respiration or nitrification (Harley *et al.*, 2006; Joint *et al.*, 2010), any general increase in OA due to changes in atmospheric composition will ultimately lift the whole baseline of ecosystems to higher acidification level with peaks above today's maxima. The natural laboratory that we chose here is defined by long-term acidification through  $\text{CO}_2$  vents that have been active for at least a century (Fabricius *et al.*, 2011). We use this system to characterize potentially adapted microbial communities to OA peak levels expected at the end of this century and beyond (700–1500  $\mu\text{atm } p\text{CO}_2$ ). Our main hypothesis is that long-term exposure to higher  $\text{CO}_2$  significantly affects the diversity of bacterial and archaeal communities. We focused on changes in local (sample) richness ( $\alpha$ -diversity) and on community turnover between

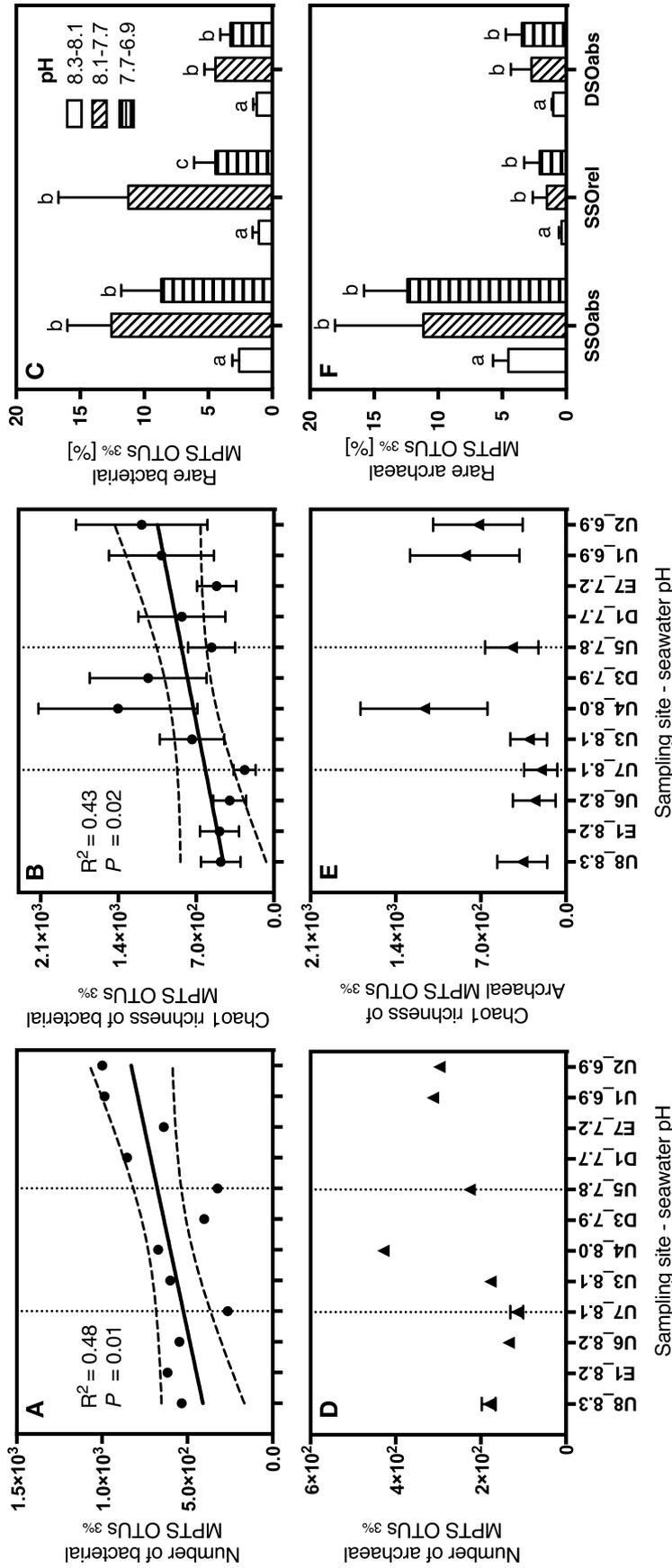
samples ( $\beta$ -diversity) while also taking into consideration effects on rare community members. Our second hypothesis is that it is possible to identify bacterial or archaeal taxa mostly affected by increased  $p\text{CO}_2$  and that such taxa may be indicators of important biogeochemical processes such as the carbon, nitrogen or sulfur cycles reacting sensitively to OA.

## Results and discussion

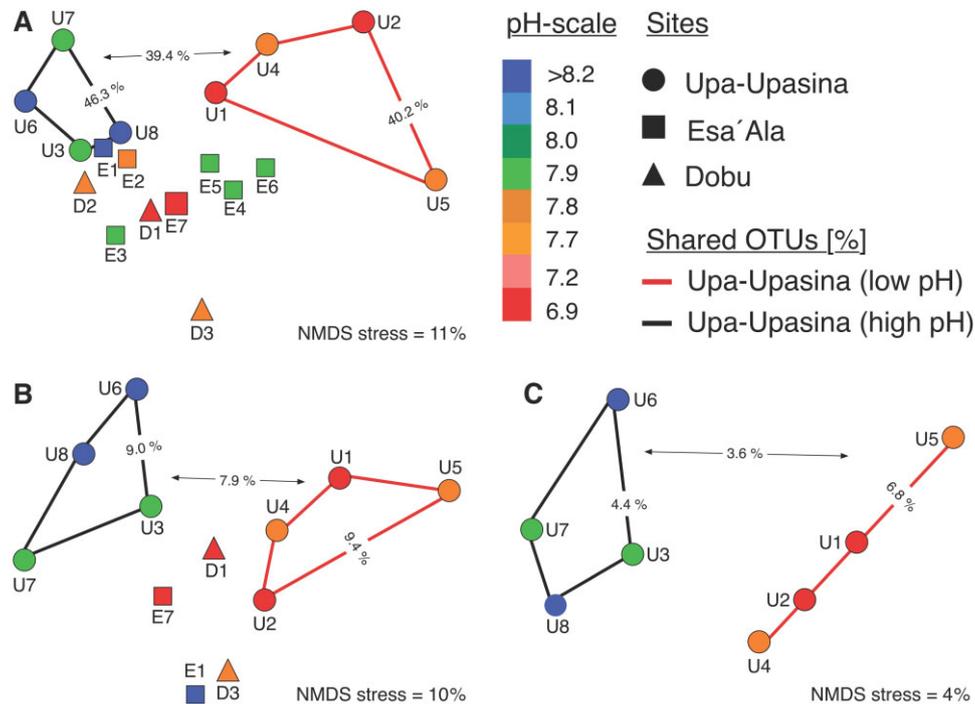
Sampling sites consisted of three geographically separated natural  $\text{CO}_2$ -seeping areas called Upa-Upasina, Esa'Ala and Dobu, and of adjacent control sites (Table S1, Fig. S1) (Fabricius *et al.*, 2011). Because we could not find any significant location-based difference in overall community structure among the three sampling areas, but mostly  $\text{CO}_2$  effects, we considered the found  $\text{CO}_2$  effects on changes in microbial diversity of general relevance.

### Diversity

*$\alpha$ -Diversity.* Both observed and estimated bacterial operational taxonomic unit (OTU) richness, as determined by 454 massively parallel tag sequencing (MPTS), increased with increasing  $p\text{CO}_2$  levels (linear regression slope coefficients = 324.7,  $P = 0.0012$ , and 105.6,  $P = 0.021$  respectively; Fig. 1, Table S2). No increase in bacterial richness was, however, detected by the community fingerprinting technique automated ribosomal intergenic spacer analysis (ARISA) (Fig. S2, Table S2), which is known to offer lower resolving power as compared with MPTS (Gobet *et al.*, 2013). For archaeal communities, a similar, yet weaker trend of increasing richness was found when comparing communities from sites with high  $p\text{CO}_2$  with those with medium and control  $p\text{CO}_2$  (Mann–Whitney two-tailed test,  $U = 0.0$ ,  $P = 0.03$ ) (Fig. 1, Table S2). Thorough denoising of MPTS data enabled us to include MPTS OTUs originating from the rare biosphere into our ecological analysis. We defined 'rare' OTUs as in (Gobet *et al.*, 2012), i.e. OTU occurring either as (i) one sequence in the whole data set (i.e. single sequence OTU absolute;  $\text{SSO}_{\text{abs}}$ ), (ii) one sequence in at least one sample (i.e. single sequence OTU relative to a sample;  $\text{SSO}_{\text{rel}}$ ), or (iii) two sequences in the whole data set (i.e. double sequence OTU absolute;  $\text{DSO}_{\text{abs}}$ ). Based on that definition, 67% and 68% of all bacterial and all archaeal MPTS OTUs, respectively, were considered rare (Table S2). At all those levels of rarity, a significantly lower percentage of rare bacterial and archaeal OTUs was found at control sites when compared with medium or high  $p\text{CO}_2$  sites. Although sharing a higher proportion of rare OTUs, medium and high  $p\text{CO}_2$  sites could not be significantly distinguished



**Fig. 1.** Observed and predicted (Chao 1 richness estimator) diversity of bacterial (A, B) and archaeal (D, E) communities at all sampling sites, sorted by increasing  $p\text{CO}_2$ . Richness estimators were calculated by subsampling at the lowest number of sequences in the whole data set ( $n = 856$  and  $918$  for bacteria and archaea respectively). The average percentages per sample of rare bacterial and archaeal MPTS OTUs<sub>3%</sub> (as compared with the total number of sequences per sample) at three  $p\text{CO}_2$  impact groups (see legend) are shown in C and F respectively. SSO<sub>abs</sub>, single sequence OTU absolute; SSO<sub>rel</sub>, single sequence OTU relative; DSO<sub>abs</sub>, double sequence OTU absolute. See text for more detail.



**Fig. 2.** Non-metric multidimensional scaling (NMDS) ordination based on Bray–Curtis dissimilarity matrices of the bacterial ARISA OTU (A), as well as bacterial (B) and archaeal (C) MPTS community tables. Symbol shape indicates site origin, and symbol colour indicates pH exposure. The black and red lines graphically highlight Upa-Upasina samples associated with low and high pH values respectively. Average percentages of numbers of shared OTUs among and between those two groups are indicated to facilitate the interpretation of community shifts.

from each other in terms of community structure (Fig. 1).

At the studied sites, Fabricius and colleagues (2011) observed a shift from coral to algal and seagrass species, and Russell and colleagues (2013) observed an increased productivity and biomass production of certain seagrass species at the  $\text{CO}_2$  seep sites as compared with control sites. Despite those changes in the communities of primary producers, sediment organic carbon (OC) and nitrogen (N) concentrations remained at the low values typically found for coral reef environments, and did not correlate with local  $\text{CO}_2$  effects (Uthicke *et al.*, 2013). Nevertheless, the different types of OC (coral versus algae and seagrass) could affect microbial communities by being more or less accessible for uptake. Additionally, because a benthic cover with fast-growing primary producers is expected to have a much higher growth rate compared with coral species, this shift could have led to an increased turnover of organic matter reaching the sediments, which would then be available as carbon and nutrient source for the reef communities. For complex bacterial communities in oligotrophic arctic sediments, a positive energy-diversity relationship has been described (Bienhold *et al.*, 2012). Thus, under the assumption that a higher turnover of carbon sources leads to more carbon-based processes in the system, the increase of bacterial and archaeal richness with  $p\text{CO}_2$  in the studied sites could

reflect changes in both energy source and availability. A different composition of carbon sources could be an additional factor favouring a richer community. Studies in terrestrial soils further support the theory that a  $p\text{CO}_2$ -induced plant growth could lead to increased bacterial biomass (Diaz *et al.*, 1993; Zak *et al.*, 1993) and more diverse communities (Berntson and Bazzaz, 1997).

**$\beta$ -Diversity.** In addition to the trend of increasing microbial richness with  $p\text{CO}_2$ , we also detected a significant shift in community composition ( $\beta$ -diversity) along the eight sites of the well-defined Upa-Upasina  $p\text{CO}_2$  gradient using MPTS data: The structure of the microbial communities that were exposed to  $p\text{CO}_2$  levels  $> 700 \mu\text{atm}$  was significantly different from the structure of communities exposed to lower  $p\text{CO}_2$  concentrations [Analysis of Similarity (ANOSIM)  $R = 0.67$ ,  $P = 0.02$ , and  $R = 0.73$ ,  $P = 0.02$ , for bacterial and archaeal communities, respectively, based on Bray–Curtis dissimilarities; Fig. 2]. For the bacterial communities, these results were highly congruent with those based on ARISA fingerprints (Fig. 2). Yet, all samples shared between 29% and 57% of OTUs based on ARISA, but only 3–16% of their OTUs based on MPTS. For all control  $p\text{CO}_2$  sites (pH 8.3–8.1), the percentage of shared bacterial OTUs was 46% and 9% based on ARISA and MPTS respectively. Similarly, for medium and high  $p\text{CO}_2$  sites (pH 8.1–6.9), the percentages of shared OTUs

were 40% and 9% for each technique respectively. When comparing all control sites with medium and high  $p\text{CO}_2$  sites, microbial communities shared only 39% of ARISA and 8% of MPTS bacterial OTUs. Archaeal communities shared between 2% and 11% of MPTS OTUs among all sites. Control sites shared 4% of archaeal MPTS OTUs, whereas sites with medium and high  $p\text{CO}_2$  shared 7% of archaeal MPTS OTUs. Control and high  $p\text{CO}_2$  sites shared only 4% of archaeal MPTS OTUs.

A multivariate analysis with all available environmental data further supported the importance of  $\text{CO}_2$  effects on changes in microbial community structure. The best redundancy analysis (RDA) model, which included the factors  $p\text{CO}_2/\text{pH}$ , carbon content and nitrogen content, explained 49% of bacterial community structure variation ( $F_{5,6} = 1.2$ ,  $P = 0.018$ , based on 1000 permutations), whereas the level of acidification (pH) alone significantly contributed with 9% ( $F_{1,10} = 1.04$ ,  $P = 0.030$ ). The model for archaea explained 67% of community structure variation, whereas the level of acidification alone was marginally significant and explained 18% ( $F_{1,6} = 1.3$ ,  $P = 0.060$ ) of the total community structure variation.

Besides community shifts directly related to environmental parameters, we also assessed possible time-related shifts. For this analysis only, we included bacterial MPTS data from Dobu sites, based on samples taken 1 year after the first sampling (referred to Dobu in 2010 and in 2011 respectively). We found the bacterial community structures of year 2011 to be significantly different from those of 2010 (ANOSIM  $R = 0.74$ ,  $P < 0.001$ ). Nevertheless, when communities from all three sites and both sampling years were analysed together, we found bacterial communities from control sites to be significantly different from those at high  $p\text{CO}_2$  sites ( $R = 0.34$ ,  $P = 0.017$ ) (Fig. S3). Communities from medium  $p\text{CO}_2$  sites could not be statistically differentiated from those of the control or of the high  $\text{CO}_2$  groups when 2010 and 2011 communities were compared.

Overall, these results indicate that high  $p\text{CO}_2$  sites not only favour a richer community, in terms of  $\alpha$ -diversity, but also select for different dominant microbial types. Although the aforementioned changes in ecosystem structure, such as increased benthic cover and higher nutrient availability, could be responsible for such a restructuring of microbial communities, there could also be  $p\text{CO}_2$  effects on specific processes via enzyme kinetics (Piontek *et al.*, 2013) or cell homeostasis (Krulwich *et al.*, 2011). It has been experimentally shown that a change in seawater chemistry (the stress factor) results in higher abundance and production of complex bacterioplankton communities (Bouvier *et al.*, 2012), and similar observations have been made in other studies (Boles *et al.*, 2004; Girvan *et al.*, 2005). At our study sites, we do not expect such microbial responses to short-term

variations in a stress factor as being typical for the analysed communities because the investigated venting areas are supposed to have been stable for several decades (Fabricius *et al.*, 2011). Ecologically speaking, the observed increase in microbial species diversity and the larger proportion of rare microbial types at high  $p\text{CO}_2$  sites could potentially have a stabilizing effect on local biological processes and ecosystem functioning: Especially, a richer community of rare types might offer a reservoir of new or interchangeable functions (Gobet *et al.*, 2012) that might be able to replace those that could be lost or weakened by environmental pressure.

#### Identification of key microbial types affected by $\text{CO}_2$ effects

**General compositional changes.** From a total of 39 535 bacterial sequences in our main data set, we taxonomically identified 4892 (out of 7443) bacterial OTUs, among which 1.5% occurred in at least 50% of all sites and two of the three sampling areas. From these 1.5%, 27% had a significant positive and 21% a significant negative linear relationship with increasing  $p\text{CO}_2$ , corresponding to 41% and 18% of bacterial sequences in the data set respectively. From a total of 21 021 archaeal sequences, we taxonomically identified 2749 (out of 3367) archaeal OTUs, among which 1.9% occurred in at least 50% of all sites in the Upa-Upasina sampling area. From these 1.9%, 15% had a significant positive and 2% a significant negative linear relationship with increasing  $p\text{CO}_2$ , corresponding to 25% and 37% of archaeal sequences in the data set respectively.

To further identify the microbial types that contributed the most to the observed changes in microbial richness and community composition at high  $p\text{CO}_2$  sites, we performed a detailed taxonomic analysis of the 16S rRNA gene libraries: From all taxonomically identified bacterial MPTS OTUs, taxonomic information could be assigned to 98.5% on the phylum, 96.7% on the class level and 80.9% on the family level. From all taxonomically identified archaeal OTUs, 96.5% could be taxonomically assigned to the phylum level, 91.3% to the class level and 48.6% to the family level. At all sites, the communities were dominated by the bacterial classes *Alphaproteobacteria*, *Deltaproteobacteria*, *Gammaproteobacteria*, *Actinobacteria*, *Flavobacteria*, *Acidobacteria* and *Planctomycetacia* (Fig. S4). The archaeal communities were dominated at all sites by the *Crenarchaeota* classes *Marine Group I*, *Thermoprotei* and *pSL12* as well as the *Euryarchaeota* classes *Thermoplasmata*, *Methanococcus* and *Halobacteria* (Fig. S5).

Among the bacterial types (here considering various levels of taxonomic resolution) that only occurred at control, medium or high  $p\text{CO}_2$  sites, none could be

conspicuously associated to a known microbial function (Table S3). One explanation could be that a  $p\text{CO}_2$  increase of approximately  $1000 \mu\text{atm}$  is not extreme enough to favour the development of completely different or endemic microbial communities when compared with local background (control) conditions, but rather leads to gradual shifts among the coexisting communities. Supporting this hypothesis, we found significant linear increases in relative sequence abundances with decreasing pH for a large number of bacterial and archaeal sequences when the  $P$  values of the partial multiple regression models were set to  $< 0.1$ .

At broad taxonomic resolution levels, we found significant linear increases in relative sequence abundance with decreasing pH for the bacterial phyla *Bacteroidetes* ( $R^2 = 0.47$ ,  $P = 0.014$ ) and *Proteobacteria* ( $R^2 = 0.35$ ,  $P = 0.042$ ), as well for the dominant archaeal *Crenarchaeota* class *Thermoprotei* ( $R^2 = 0.43$ ,  $P = 0.078$ ) (Table S4). An overall significant decrease of sequence abundances with decreasing pH was found for the bacterial phyla *Nitrospirae* ( $R^2 = 0.38$ ,  $P = 0.032$ ) and *Actinobacteria* ( $R^2 = 0.28$ ,  $P = 0.079$ ) as well as for the *Crenarchaeota* class *Marine Group I* ( $R^2 = 0.49$ ,  $P = 0.052$ ; Fig. 3) (Table S5).

*Taxa potentially associated with the nitrogen cycle.* *Marine Group I* are suspected to be among the most abundant marine microorganisms (DeLong, 1992) and made up to 37% of all archaeal sequences in our data set. Nevertheless, one prominent OTU, whose representative sequence was identified as *Nitrosopumilus maritimus* [99% nucleotide (nt) identity], made up to 13% of all archaeal sequences and increased in relative sequence abundance with seawater acidification ( $R^2 = 0.51$ ,  $P = 0.046$ ; Fig. 3; Table S5). As a member of the newly defined phylum *Thaumarchaeota*, which consists of a major group of ammonia-oxidizing archaea (AOA), previously assigned to the diverse *Marine Group I* (Könneke *et al.*, 2005), *N. maritimus* is known for its high affinity to oxygen and for growing chemoautotrophically on ammonia with inorganic carbon as sole carbon source (Walker *et al.*, 2010). Concomitantly, the only sequences of ammonia-oxidizing bacteria (AOB) identified in our data significantly decreased with increasing seawater acidification and were identified as *Gammaproteobacteria Nitrosococcus* sp. ( $R^2 = 0.32$ ,  $P = 0.057$ ; Fig. 3; Table S4), representing 1% of all bacterial sequences.

It is known that many AOA, in contrast to AOB, seem to be adapted to thrive even under oligotrophic ammonia concentrations, with *N. maritimus* even showing the highest substrate affinity to ammonia among all microorganisms known to date (Martens-Habbena *et al.*, 2009). Another very distinctive property of many *Thaumarchaeota*, which is of special relevance to OA, is

their ability to perform nitrification at neutral or lower pH: Lehtovirta-Morley and colleagues (2011) even reported an acidophilic nitrifying *Thaumarchaeota*, isolated from soil, with an optimum pH  $< 5$ . Despite the ratios of ammonia oxidation to nitrite being in the range of known AOB, it has therefore been suggested that AOA might use a different nitrification mechanism (Martens-Habbena *et al.*, 2009; Walker *et al.*, 2010; Stahl and de la Torre, 2012).

Although most strains of nitrifying bacteria are known to be pH sensitive with an optimal growth range between pH 7 and pH 8 (Tarre *et al.*, 2004), there is evidence for the existence of acid-tolerant AOB communities (De Boer and Kowalchuk, 2001; Gieseke *et al.*, 2006). Acid tolerance does not seem to be connected to any uncommon AOB taxa, but clearly requires physiological adaptations, such as ammonium ion transporters (Gieseke *et al.*, 2006) because AOB require  $\text{NH}_3$  for their ammonia mono-oxygenase. Because  $\text{NH}_3$  is decreasing 10-fold with any 1-unit reduction in pH through protonation to  $\text{NH}_4^+$ , non-acid-tolerant communities may become substrate limited in acidified seawater (Stahl and de la Torre, 2012). It is not yet clear how common acid-tolerant AOB communities are in marine coastal sediments. Our observation of a decrease in AOB sequences therefore tends to support a previous study that reported a decrease in bacterial ammonia oxidation rates as a common response to acidification in marine environments (Beman *et al.*, 2011), while keeping in mind that changes in sequence relative abundance do not necessarily translate directly to changes in process rates.

Although AOA seem to dominate ammonia-depleted regions of the ocean, there are many environments, such as coastal ecosystems, where they are outcompeted by substrate competition with AOB and phytoplankton for ammonia (Martens-Habbena *et al.*, 2009; Stahl and de la Torre, 2012). Although we could not find any significant change in sequence abundances of major nitrogen-fixing bacteria or archaea in the sediment, Hutchins and colleagues (2009) reported evidence for elevated nitrogen fixation rates in the water column, based on a fertilizing effect of  $\text{CO}_2$  as well as the range expansion of temperature sensitive nitrogen-fixing organisms, in a warmer ocean. Potentially higher availability of ammonium, from nitrogen fixation in the water column, which was not measured in this study, might offer an advantage to AOB, whereas AOA, which generally seem better adapted to oligotrophic ammonium levels (Stahl and de la Torre, 2012), could profit from lower pH levels and from the depletion of AOB in an acidified future ocean. In a short-term incubation experiment with coastal arctic sediment, Tait and colleagues (2014) could show that 2 weeks of increased  $p\text{CO}_2$  exposure at  $760 \mu\text{atm}$  leads to a decreased transcription of bacterial ammonia-oxidizing genes (*amoA*), whereas archaeal *amoA* transcript

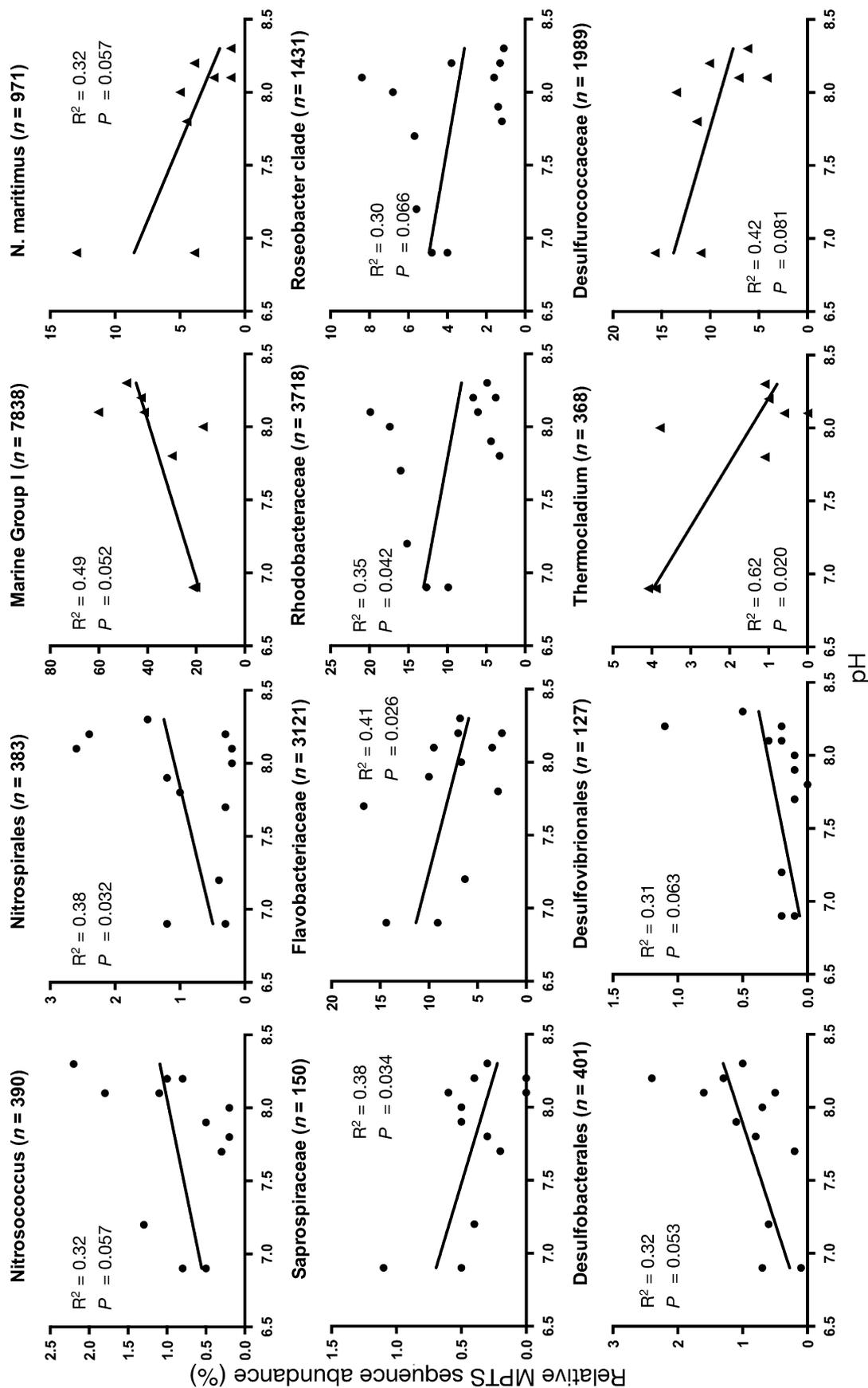


Fig. 3. Significant linear relationships of relative MPTS sequence abundances (y-axes) with pH (x-axes) of selected bacterial (circles) and archaeal (triangles) taxa. The taxonomic group and the total number of sequences in the data set ( $n$ ) are displayed on top of each graph. For the statistical analysis, the  $P$  values of the partial multiple regression models were set to  $< 0.1$ .

numbers seemed not to be affected. At such acidification levels, Gazeau and colleagues (2014) could not detect any effect on the simultaneously measured sediment nitrogen fluxes within the limited time frame of this experiment, suggesting that AOA could, up to a certain point, compensate for putatively lower AOB activity. Such an activity shift from AOB to AOA could also explain the finding by Kitidis and colleagues (2011), who detected no effect on ammonia oxidation in coastal sediments with a history of long-term exposure to increased CO<sub>2</sub> from vents off the island of Ischia in the Mediterranean Sea.

Although in our data set, the increased occurrence of *N. maritimus* sequences with acidification was accompanied by a decrease in AOB sequences, it is interesting to note that sequences of all identified *Nitrospirales*, which made up 49% of all nitrite-oxidizing bacteria (NOB) identified in our data set, decreased with lower pH values ( $R^2 = 0.38$ ,  $P = 0.032$ ; Fig. 3; Table S4). This could suggest that ammonia oxidation rates might in fact be lower in communities that have an increased proportion in AOA, and this hypothesis should be tested in future work. Nevertheless, the quantitative aspect of the trends reported in our study should be taken with caution as it is the case with any polymerase chain reaction (PCR)-based methods, which do not represent true changes in organism abundances.

*Taxa potentially related to the sulfur cycle.* It is not expected that seawater acidification directly affects sulfur or sulfate reduction in sediments (Koschorreck, 2008), but among the sulfate-reducing bacteria, we found sequences of two obligate anaerobic deltaproteobacterial groups, the acetate-oxidizing *Desulfobacterales* and the non-acetate-oxidizing *Desulfovibrionales* with a significant negative linear relationship with increasing seawater acidification ( $R^2 = 0.32$ ,  $P = 0.053$ ; and  $R^2 = 0.31$ ,  $P = 0.063$  respectively; Fig. 3; Table S4). Together, they represent 14% of all bacterial sequences in the data set that could be identified as potential sulfate reducers. In contrast to bacterial sulfate reducers, we found a linear increase with acidification for sequences of the sulfate-reducing *Crenarchaeota* family *Thermocladium* ( $R^2 = 0.62$ ,  $P = 0.020$ ; Fig. 3; Table S5), which represented 22% of all putative sulfate-reducing archaea in our data set as well as the sulfur-reducing *Crenarchaeota* order *Desulfurococcaceae* ( $R^2 = 0.42$ ,  $P = 0.081$ ; Fig. 3; Table S5). Their OTUs were distantly related to *Thermocladium modestius* (74% nt identity), which has been shown to grow optimally at pH 4.0 and *Ignicoccus pacificus* (74% nt identity) that can grow between pH 4.5–7.0 respectively (Itoh *et al.*, 1998; Huber *et al.*, 2000).

*Taxa potentially related to extreme conditions.* *Thermocladium* species and *Desulfurococcaceae* species

seem to be limited to extreme temperatures from 60 to 100°C (Madigan *et al.*, 1997; Huber *et al.*, 2000). Besides *Thermocladium*, we found sequences of other putatively extremophilic microorganisms with positive linear relationships to decreasing pH, such as sequences of the *Euryarchaeota* class of *Thermoplasmatales*, which contains three thermophilic and extremely acidophilic genera, including the most acidophilic of all known organisms with a favoured growth range at pH 0.5–4.0 (Cowan, 2000). We found bacterial sequences associated with the genus *Rhodothermus*, which is currently assigned to the *Rhodothermaceae* of the phylum *Bacteroidetes*, to significantly increase in abundance with increasing CO<sub>2</sub> effect ( $R^2 = 0.30$ ,  $P = 0.065$ ; Table S4). Species described so far grow at a pH range of 6.0–8.0 and temperatures range of 55–80°C (Marteinsson *et al.*, 2010). Also, an increase in sequences of the order *Flavobacteriaceae*, which is known to have very limited tolerance to elevated temperatures above 50°C and which is one of the most dominant bacterial groups found at our high pCO<sub>2</sub> sites, questions the occurrence of an active extremophilic community in the sampled sediments. Because the seeping is of volcanic origin, one explanation could be that extremophilic members actually grow in the deeper hot subsurface and are transported to the surface through CO<sub>2</sub> seepages. As such, they might be considered as contaminants of the surface sedimentary communities.

*Taxa related to the carbon cycle.* Among the dominant bacterial phyla, *Bacteroidetes* (10% of all bacterial sequences), we found sequences associated with the order *Saprospiraceae*, which significantly increased with increasing pCO<sub>2</sub> ( $R^2 = 0.38$ ,  $P = 0.034$ ; Fig. 3; Table S4). *Saprospiraceae* are well known as major heterotrophic consumers of plant biomass and are common in marine littoral sand and coastal zones in various locations worldwide (Delk and Dekker, 1972; Saw *et al.*, 2012). Sequences of the dominant family *Flavobacteriaceae* (8% of all bacterial sequences) significantly increased with seawater acidification ( $R^2 = 0.41$ ,  $P = 0.026$ ; Fig. 3; Table S4), with the most dominant OTUs being closely related to *Aquimarina* species, especially one prominent OTU ( $R^2 = 0.53$ ,  $P = 0.007$ ; Table S4), which made up to 14% of all *Flavobacteriaceae* sequences and was closely related to *Aquimarina mytili* (95% nt identity). This species was recently isolated and described to grow optimally at pH 7.0 and 25–30°C (Park *et al.*, 2012). Another prominent *Flavobacteriaceae* OTU ( $R^2 = 0.33$ ,  $P = 0.051$ , representing 26% of all *Flavobacteriaceae*; Table S4) was found to be closely related to *Actibacter sediminis* (96% nt identity) with an optimum growth at at pH 6.0 and 37°C (Kim *et al.*, 2008). Generally, most members of the *Flavobacteriaceae* are all widely distributed in oxic marine environments and are aerobic chemoheterotrophs.

Typically, they abundantly occur in sediments when sufficient oxygen is available, for instance within surface mixed layers with oxygenating influence by bioturbation or oceanic currents (Bowman, 2006). Similar to *Saprospiraceae*, *Flavobacteriaceae* play a prominent role in the degradation of complex polymeric substrates in marine environments and are often associated with the occurrence of large amounts of plant detrital biomass (Bowman, 2006; Klippel *et al.*, 2011). The higher cover of algae and seagrass at the high  $p\text{CO}_2$  sites might therefore be an even more relevant reason for the observed increase in both *Flavobacteriaceae* and *Saprospiraceae* sequences.

Another major group of bacteria that significantly increased in sequence abundance with  $p\text{CO}_2$  was the *Alphaproteobacteria* order *Rhodobacteraceae* ( $R^2 = 0.35$ ,  $P = 0.042$ , representing 9% of all bacterial sequences; Fig. 3; Table S4). These are known to include acidophilic types, with a growth optimum at pH 6.0 (Madigan *et al.*, 1997). But among the *Rhodobacteraceae* bacteria, the group of the so-called *Roseobacter* clade seemed to most prominently profit from low pH conditions (linear increase  $R^2 = 0.30$ ,  $P = 0.066$ ; Fig. 3; Table S4), making up to 39% of all *Rhodobacteraceae* sequences. Two very prominent OTUs with positive linear relationships to decreasing pH were closely related to *Phaeobacter daeponensis* (100% nt identity), which was isolated from a tidal flat in the Yellow Sea, Korea, and to *Phaeobacter gallaeciensis* (92% nt identity), which was isolated from the German Wadden Sea. Both species have been reported to grow optimally at seawater pH < 7.0–8.0 and have a chemoheterotrophic and obligate aerobic metabolism (Martens *et al.*, 2006; Yoon *et al.*, 2007). Generally, bacteria of this group have been isolated from habitats all over the world and are known to be one of the most abundant groups in marine environments (Martens *et al.*, 2006).

Many *Flavobacteriaceae*, *Saprospiraceae* and *Roseobacter* types are often found associated with marine algae, seagrasses and coastal biofilms (Buchan *et al.*, 2005). Interestingly, Teira and colleagues (2012) observed no direct effect on cell numbers or biomass of *Rhodobacteraceae* strain MED165 and *Flavobacteriaceae* strain MED217 when experimentally treated with increased  $p\text{CO}_2$  alone. This supports our hypothesis that those bacterial groups could profit from a  $p\text{CO}_2$ -induced increase in algal and seagrass biomass instead of being directly influenced by the seawater chemistry.  $\text{CO}_2$  effects of increased seawater acidification on *Alphaproteobacteria* and *Rhodobacteraceae* as well as *Bacteroidetes* and *Flavobacteriaceae* have been reported before: At Upa-Upasina, also one of our study sites, Morrow and colleagues (2014) observed that increasing  $\text{CO}_2$  in the seawater was accompanied by an increase in

relative MPTS sequence abundance of *Flavobacteria* associated with the hard coral *Porites cylindrica*. Also Vega Thurber and colleagues (2009) experimentally showed that numbers of *Flavobacteriaceae* on coral surfaces increased under low pH conditions, whereas Meron and colleagues (2010) reported a positive effect of acidified seawater on the occurrence of *Rhodobacteraceae* on coral tissue. Nevertheless, because many *Flavobacteriaceae* and *Rhodobacteraceae* are also associated with coral diseases (Cooney *et al.*, 2002; Meron *et al.*, 2010), the reason for an increase in abundances could, in those cases, be related to the corals themselves, which may be stressed by acidification and therefore more prone to infections.

In biofilms developing on glass slides and exposed to acidified seawater from the Great Barrier Reef, some *Flavobacteriaceae* were reported to increase in relative abundances with decreasing pH, whereas, in contrast to our results, members of the *Roseobacter* clade showed a decreasing trend with rising  $p\text{CO}_2$  (Witt *et al.*, 2011). In addition, Krause and colleagues (2012) experimentally showed that the community composition of microbial batch cultures, originating from the North Sea water column, significantly changed under the influence of small reductions in seawater pH and was accompanied by pH-dependent dissimilarities mostly caused by changes in *Flavobacteriaceae* and by a higher proportion of *Rhodobacteraceae* at ambient seawater  $p\text{CO}_2$ .

For some bacterial and archaeal taxa, which generally showed significant linear relationships with decreasing pH, we found sequence occurrences of subgroups, which could better be described by quadratic relationships, which suggests modal distributions rather than linear changes along the pH gradient. Nevertheless, with the exception of sequences related to unidentified *Marine Group I* OTUs, quadratic relationships of sequence occurrences with pH were scarce in our data set and seemed to represent rather complex relationships, for which it might be too speculative to discuss functional ecological patterns apart from possible niche differentiations. Nevertheless, we provide the results based on quadratic relationships and those related to microbial taxa without obvious ecological functions in the Supporting Information section (Text S1 and Tables S6 and S7).

## Conclusions

Our results strongly support the hypothesis that the nitrogen cycle might be directly affected by rising oceanic  $p\text{CO}_2$ , with possible disadvantages for AOB and advantages for better adapted AOA. At our study sites,  $\text{CO}_2$ -induced seawater acidification leads to a gradual increase in the cover of benthic primary producers, namely algae and seagrasses. The expected increase in biomass turno-

ver may explain the observed increase in organic matter degrading bacterial groups from the *Flavobacteriaceae*, *Roseobacter* clade or *Saprospiraceae*. Besides these changes in taxonomic groups, we clearly evidenced an increase in bacterial and archaeal richness and an increase in rare OTUs along the acidification gradient. Even after factoring out the temporal shift in communities between two sampling years, we still observed a strong structuring effect of increased  $p\text{CO}_2$  on bacterial and archaeal communities. Therefore, our study clearly reveals the sensitivity of complex natural bacterial and archaeal communities to seawater acidification, as expected for our future oceans, and identified the microbial components that are most likely to be affected.

## Experimental procedures

### *Samples and sites*

The three natural  $\text{CO}_2$ -seeping areas, called Upa-Upasina, Esa'Ala and Dobu, are located around the D'Entrecasteaux Islands in the Milne Bay Province of Papua New Guinea (Fig. S1) and were first described by Fabricius and colleagues (2011). The gas composition at the studied sites has been measured by Fabricius and colleagues (2011). At all three venting areas, the gas consists of over 99%  $\text{CO}_2$ , with traces of  $\text{O}_2$ ,  $\text{N}_2$  and  $\text{CH}_4$ , whereas no  $\text{C}_2\text{H}_2$  or  $\text{N}_2\text{O}$  could be detected. The gas emitted at Upa-Upasina and Esa'Ala is  $\text{H}_2\text{S}$ -free, whereas the gas at Dobu contains about 163 ppm  $\text{H}_2\text{S}$ . Arsenic and 10 heavy metals (V, Cr, Co, Ni, Cu, Ga, Mo, Cd, Pb and U) were measured in the seawater above the benthos at all control and seep areas by Uthicke and colleagues (2013). They found no significant difference between seep and non-seep sites, and all concentrations were in ranges expected for pristine seawater. All seeps investigated in this study are assumed to have existed for at least seven decades (Fabricius *et al.*, 2011).

Our study was based on 18 samples taken in 2010 and 6 samples taken in 2011, of the upper three centimetres of the oxic zone of sandy sediment at a water depth of 1–4 m (Table S1). All samples were immediately frozen and stored at  $-20^\circ\text{C}$  until analysis. Control sites consisted of five sediment-sampling sites in 2010 and two in 2011, which were located at proximity of the seeps with very similar environmental settings, but not directly affected by  $\text{CO}_2$  emissions. Their seawater pH levels ranged from 8.3 to 8.1 (NBS scale), corresponding to typical ambient  $p\text{CO}_2$  concentrations of 360–400  $\mu\text{atm}$ . The 13 sediment-sampling sites from 2010 with increased seawater acidification levels were sampled to represent a gradient, ranging from pH 8.0 to pH 6.9 ( $p\text{CO}_2$  concentrations of 620–1500  $\mu\text{atm}$ ). The most continuous pH gradient was found at the site Upa-Upasina, from which eight samples were taken in 2010, including two control areas at opposite sites from the venting area along the coast. In 2011, two additional samples were taken from a  $\text{CO}_2$ -seeping site near Dobu with an average pH of 7.7, as well as two samples from a close-by venting area with an average pH of 6.8, and two from the Dobu control site (pH 8.2). Although the samples from 2010 represent the main data set for this study, the six samples from 2011 were used to get an estimate of commu-

nity shift over time. The pH values in the overlaying water and in the upper sediment layer were measured during the field expeditions in 2010 and 2011, when the samples for this study were taken.

In terms of benthic cover, Fabricius and colleagues (2011) observed that with decreasing seawater pH, structurally complex coral communities gradually shift to assemblages dominated by slow-growing, long-lived and structurally simple *Porites*. Despite this change in community composition, coral cover remained stable in the range of pH 8.1–7.8, whereas no reef development was found at sites with  $> 1000 \mu\text{atm } p\text{CO}_2$ . For plant and algal cover, a loss of coralline species with decreasing pH was accompanied by an increase in seagrass and non-calcareous macroalgae (Fabricius *et al.*, 2011; Russell *et al.*, 2013). In surface sediments, concentrations of OC, nitrogen and siliceous spicules did not change along the gradient, whereas inorganic carbon, for instance in form of the remains of calcareous species, was only found in relevant amounts at sites with  $p\text{CO}_2 < 1000 \mu\text{atm}$  (Uthicke *et al.*, 2013).

### *DNA extraction and microbial community fingerprinting*

Total community DNA from each sample was isolated from 1 g well-mixed sediment using UltraClean Soil DNA Isolation kits (MoBio Laboratories, Carlsbad, CA, USA). Two aliquots with a final volume of 50  $\mu\text{l}$  of Tris-EDTA buffer were stored at  $-20^\circ\text{C}$  until analysis. DNA concentrations were determined spectrophotometrically using a NanoQuant infinite M2000 (Tecan Group, Männedorf, Switzerland). The DNA of all 18 sediment samples from 2010 was analysed via the molecular fingerprinting technique ARISA (Fisher and Triplett, 1999) using a triplicate PCR approach. For each PCR, the amount of DNA was adjusted to 25 ng. PCR amplification, analysis of amplified fragments via capillary electrophoresis and subsequent binning into OTUs were done as described previously (Ramette, 2009). An OTU was considered present in a sample only if it appeared in at least two out of the three PCR replicates for that sample.

### *MPTS*

The combination of the molecular techniques ARISA and MPTS, used in this study, has been successfully applied before (Bienhold *et al.*, 2012; Gobet *et al.*, 2013; Jacob *et al.*, 2013), because it allows an extensive assessment of overall  $\alpha$ - and  $\beta$ -diversity, while leaving the opportunity of targeting specific samples via MPTS for a deeper diversity analysis and for taxonomic identification of key microbial indicators. Samples for which complete and comparable environmental parameters were available were chosen for MPTS. The sequencing was done via the Roche 454 FLX+ system (Roche, Basel, Switzerland) at the Research and Testing Laboratory (Lubbock, Texas, USA). For bacterial communities, DNA of 12 samples from 2010 and all 6 samples from 2011 were amplified with primers targeting the bacterial V1-V3 region (Gray28F 5'-TTTGATCNTGGCTCAG-3'). For the archaeal community, eight samples from 2010 from site Upa-Upasina were amplified with primers targeting the archaeal V6 region (Arch349F 5'-GYGCASCAGKCGMG AAW-3'). Sequences were deposited with the European

Nucleotide Archive under study accession number PRJEB7705 and sample accession numbers ERS580491-ERS580498 for archaeal samples U1 to U8, respectively, and sample accession numbers ERS580473-ERS580490 for bacterial samples U1-U8, E1, E7, D1, D3, CST, CSB, BST, BSB, ST and SB respectively.

Denosing (i.e. pyrosequencing noise removal), trimming, error correction (chimera and PCR errors) of the MPTS sequences were implemented with the *mothur* pipeline (version 1.23.1; Schloss *et al.*, 2009). For alignment and taxonomic assignment of the bacterial and archaeal sequences, SILVA reference databases (both, version 115) provided with the *mothur* pipeline were used. MPTS OTUs were defined at the 3% nt difference level. OTU richness was estimated with the Chao1 richness estimator, which takes into consideration the number of rare sequences while correcting for the estimated undiscovered species being present in a sample (Chao, 1984). Chao1 richness was only calculated after re-sampling every sample to the lowest number of sequences found in any sample. The inverse Simpson's index (Simpson, 1949) was used to measure the evenness of the diversity in each sample as the index is maximized when all OTUs are equally represented. For all bacterial and archaeal OTUs significantly changing with the  $p\text{CO}_2$  gradient, we performed a systematic search for described microbial nucleotide sequences via the BASIC LOCAL ALIGNMENT SEARCH TOOL (BLAST; standard nucleotide data base for bacterial and archaeal 16S rRNA genes; Altschul *et al.*, 1990) using the best representative sequences for each OTU, followed by filtering of the results for sequences matching with > 90% nt identity, > 98% query coverage and E-value <  $10^{-6}$ .

### Statistical analysis

The multivariate statistical analyses of all molecular data from ARISA and MPTS, along with the measured environmental parameters, were performed with the R statistical language (version 2.15.0; R Core Team, 2013, URL: <http://www.R-project.org>), using the R packages 'MASS' (Venables and Ripley, 2002) and 'vegan' (Oksanen *et al.*, 2013) to perform non-metric multidimensional scaling, one-way ANOSIM, stepwise model selection, canonical RDA and associated permutation tests (reviewed in Ramette, 2007). For the single OTU analysis, changes in the sequence abundance of an OTU were considered to be significant when the *P* values of the partial multiple regression model (1000 permutations) were < 0.1, after false discovery rate correction (Benjamini and Hochberg, 1995). To quantify the specific  $p\text{CO}_2$  effects on community variation while disentangling the effects of other measured parameters, we performed RDA with models including the environmental parameters  $p\text{CO}_2/\text{pH}$ , total carbon content, nitrogen content, as well as geographic distances between the sampling sites (Table S1, partly published by Uthicke *et al.*, 2013). Scripts of the MULTICOLA software (version 1.5, available at URL: [http://www.mpi-bremen.de/Software\\_2.html](http://www.mpi-bremen.de/Software_2.html); Gobet *et al.*, 2010) were used to assess the effects of rare MPTS OTUs on the reported ecological interpretations. For some statistical analysis, sites were grouped into the three categories (control, medium and high  $p\text{CO}_2$ ), corresponding to the pH ranges of 8.3–8.1, 8.1–7.7 and 7.7–6.9 respectively. Statistical tests to compare

models of linear and quadratic relationships of relative MPTS sequence abundances with environmental parameters were performed as described in Makarenkov and Legendre (2002). Significant linear relationships of taxa and OTUs were only considered if there was no significantly better quadratic model found.

### Acknowledgements

We thank the Australian Institute of Marine Science (AIMS) for funding the field work, the owners of the Upa-Upasina, Dobu and Esa'Ala reefs for allowing us to study their reefs, and the crew and Captain of the Chertan (Milton Bay Enterprises). We also thank the Dobu RLLG, the Milne Bay Province Research Committee, and the Department of Environment and Conservation of Papua New Guinea for delivering permits and for the logistic support. The study was supported by the Deutsche Forschungsgemeinschaft and the Federal Ministry of Education and Research (BMBF) in the framework of the BIOACID project, and by the Max Planck Society.

### References

- Altschul, S.F., Gish, W., Miller, W., Myers, E.W., and Lipman, D.J. (1990) Basic local alignment search tool. *J Mol Biol* **215**: 403–410.
- Andersson, A.J., Bates, N.R., and Mackenzie, F.T. (2007) Dissolution of carbonate sediments under rising  $p\text{CO}_2$  and ocean acidification: observations from Devil's Hole, Bermuda. *Aquat Geochem* **13**: 237–264. doi:10.1007/s10498-007-9018-8.
- Beman, J.M., Chow, C.E., King, A.L., Feng, Y., Fuhrman, J.A., Andersson, A., *et al.* (2011) Global declines in oceanic nitrification rates as a consequence of ocean acidification. *Proc Natl Acad Sci USA* **108**: 208–213. doi:10.1073/pnas.1011053108.
- Benjamini, Y., and Hochberg, Y. (1995) Controlling the false discovery rate: a practical and powerful approach to multiple testing. *J Roy Stat Soc B* **57**: 289–300.
- Berntson, G.M., and Bazzaz, F.A. (1997) Nitrogen cycling in microcosms of yellow birch exposed to elevated  $\text{CO}_2$ : simultaneous positive and negative below-ground feedbacks. *Glob Change Biol* **3**: 247–258.
- Bienhold, C., Boetius, A., and Ramette, A. (2012) The energy-diversity relationship of complex bacterial communities in Arctic deep-sea sediments. *ISME J* **6**: 724–732. doi:10.1038/ismej.2011.140.
- Böer, S.I., Hedtkamp, S.I.C., van Beusekom, J.E.E., Fuhrman, J.A., Boetius, A., and Ramette, A. (2009) Time- and sediment depth-related variations in bacterial diversity and community structure in subtidal sands. *ISME J* **3**: 780–791. doi:10.1038/ismej.2009.29.
- Boles, B.R., Thoendel, M., and Singh, P.K. (2004) Self-generated diversity produces insurance effects in biofilm communities. *Proc Natl Acad Sci USA* **101**: 16630–16635.
- Bouvier, T., Venail, P., Pommier, T., Bouvier, C., Barbera, C., and Mouquet, N. (2012) Contrasted effects of diversity and immigration on ecological insurance in marine bacterioplankton communities. *PLoS ONE* **7**: e37620. doi:10.1371/journal.pone.0037620.

- Bowman, J.P. (2006) The Marine Clade of the Family Flavobacteriaceae: The Genera *Aequorivita*, *Arenibacter*, *Cellulophaga*, *Croceibacter*, *Formosa*, *Gelidibacter*, *Gillisia*, *Maribacter*, *Mesonina*, *Muricauda*, *Polaribacter*, *Psychroflexus*, *Psychroserpens*, *Robiginitalea*, *Salegentibacter*, *Tenacibaculum*, *Ulvibacter*, *Vitellibacter* and *Zobellia*. In *The Prokaryotes*. Dworkin, M., Falkow, S., Rosenberg, E., Schleifer, K., and Stackebrandt, E. (eds). New York, NY, USA: Springer New York, pp. 677–694.
- Buchan, A., González, J.M., and Moran, M.A. (2005) Overview of the marine *Roseobacter* lineage. *Appl Environ Microbiol* **71**: 5665–5677. doi:10.1128/AEM.71.10.5665-5677.2005.
- Chao, A. (1984) Nonparametric estimation of the number of classes in a population. *Scand J Stat* **11**: 265–270.
- Cooney, R.P., Pantos, O., Le Tissier, M.D., Barer, M.R., O'Donnell, A.G., and Bythell, J.C. (2002) Characterization of the bacterial consortium associated with black band disease in coral using molecular microbiological techniques. *Environ Microbiol* **4**: 401–413.
- Cowan, D. (2000) Use your neighbour's genes. *Nature* **407**: 466–467. doi:10.1038/35035195.
- De Boer, W., and Kowalchuk, G.A. (2001) Nitrification in acid soils: micro-organisms and mechanisms. *Soil Biol Biochem* **33**: 853–866.
- Decho, A.W. (2000) Microbial biofilms in intertidal systems: an overview. *Cont Shelf Res* **20**: 1257–1273.
- Delk, A.S., and Dekker, C.A. (1972) Characterization of rhabdosomes of *Saprospira grandis*. *J Mol Biol* **64**: 287–295.
- DeLong, E.F. (1992) Archaea in coastal marine environments. *Proc Natl Acad Sci USA* **89**: 5685–5689.
- Diaz, S., Grime, J.P., Harris, J., and McPherson, E. (1993) Evidence of a feedback mechanism limiting plant response to elevated carbon dioxide. *Nature* **364**: 616–617.
- Fabricius, K.E., Langdon, C., Uthicke, S., Humphrey, C., Noonan, S., Death, G., *et al.* (2011) Losers and winners in coral reefs acclimatized to elevated carbon dioxide concentrations. *Nat Clim Change* **1**: 165–169. doi:10.1038/NCLIMATE1122.
- Fisher, M.M., and Triplett, E.W. (1999) Automated approach for ribosomal intergenic spacer analysis of microbial diversity and its application to freshwater bacterial communities. *Appl Environ Microbiol* **65**: 4630–4636.
- Gazeau, F., van Rijswijk, P., Pozzato, L., and Middelburg, J.J. (2014) Impacts of ocean acidification on sediment processes in shallow waters of the Arctic Ocean. *PLoS ONE* **9**: e94068. doi:10.1371/journal.pone.0094068.
- Gieseke, A., Tarre, S., Green, M., and de Beer, D. (2006) Nitrification in a biofilm at low pH values: role of in situ microenvironments and acid tolerance. *Appl Environ Microbiol* **72**: 4283–4292. doi:10.1128/AEM.00241-06.
- Girvan, M.S., Campbell, C.D., Killham, K., Prosser, J.I., and Glover, L.A. (2005) Bacterial diversity promotes community stability and functional resilience after perturbation. *Environ Microbiol* **7**: 301–313.
- Gobet, A., Quince, C., and Ramette, A. (2010) Multivariate Cutoff Level Analysis (MultiCoLA) of large community data sets. *Nucleic Acids Res* **38**: e155. doi:10.1093/nar/gkq545.
- Gobet, A., Böer, S.I., Huse, S.M., van Beusekom, J.E., Quince, C., Sogin, M.L., *et al.* (2012) Diversity and dynamics of rare and of resident bacterial populations in coastal sands. *ISME J* **6**: 542–553. doi:10.1038/ismej.2011.132.
- Gobet, A., Boetius, A., and Ramette, A. (2013) Ecological coherence of diversity patterns derived from classical fingerprinting and Next Generation Sequencing techniques. *Environ Microbiol* **16**: 2672–2681. doi:10.1111/1462-2920.12308.
- Hall-Spencer, J.M., Rodolfo-Metalpa, R., Martin, S., Ransome, E., Fine, M., Turner, S.M., *et al.* (2008) Volcanic carbon dioxide vents show ecosystem effects of ocean acidification. *Nature*. doi:10.1038/nature07051.
- Harley, C.D., Randall Hughes, A., Hultgren, K.M., Miner, B.G., Sorte, C.J., Thornber, C.S., *et al.* (2006) The impacts of climate change in coastal marine systems. *Ecol Lett* **9**: 228–241. doi:10.1111/j.1461-0248.2005.00871.x.
- Hewson, I., Jacobson Meyers, M.E., and Fuhrman, J.A. (2007) Diversity and biogeography of bacterial assemblages in surface sediments across the San Pedro Basin, Southern California Borderlands. *Environ Microbiol* **9**: 923–933. doi:10.1111/j.1462-2920.2006.01214.x.
- Huber, H., Burggraf, S., Mayer, T., Wyschkony, I., Rachel, R., and Stetter, K.O. (2000) *Ignicoccus* gen. nov., a novel genus of hyperthermophilic, chemolithoautotrophic Archaea, represented by two new species, *Ignicoccus islandicus* sp nov and *Ignicoccus pacificus* sp nov. and *Ignicoccus pacificus* sp. nov. *Int J Syst Evol Microbiol* **50** (Part 6): 2093–2100.
- Hutchins, D.A., Mulholland, M.R., and Fu, F. (2009) Nutrient cycles and Marine Microbes in a CO<sub>2</sub>-Enriched Ocean. *Oceanography* **22**: 128–145.
- Itoh, T., Suzuki, K., and Nakase, T. (1998) *Thermocladium modestius* gen. nov., sp. nov., a new genus of rod-shaped, extremely thermophilic crenarchaeote. *Int J Syst Bacteriol* **48** (Part 3): 879–887.
- Jacob, M., Soltwedel, T., Boetius, A., and Ramette, A. (2013) Biogeography of Deep-sea benthic bacteria at regional scale (LTER HAUSGARTEN, Fram Strait, Arctic). *PLoS ONE* **8**: e72779. doi:10.1371/journal.pone.0072779.
- Joint, I., Doney, S.C., and Karl, D.M. (2010) Will ocean acidification affect marine microbes? *ISME J* **5**: 1–7. doi:10.1038/ismej.2010.79.
- Kim, J.H., Kim, K.Y., Hahm, Y.T., Kim, B.S., Chun, J., and Cha, C.J. (2008) *Actibacter sediminis* gen. nov., sp. nov., a marine bacterium of the family Flavobacteriaceae isolated from tidal flat sediment. *Int J Syst Evol Microbiol* **58**: 139–143. doi:10.1099/ijs.0.65346-0.
- Kitidis, V., Laverock, B., McNeill, L.C., Beesley, A., Cummings, D., Tait, K., *et al.* (2011) Impact of ocean acidification on benthic and water column ammonia oxidation. *Geophys Res Lett* **38**: L21603. doi:10.1029/2011GL049095.
- Klippel, B., Lochner, A., Bruce, D.C., Davenport, K.W., Detter, C., Goodwin, L.A., *et al.* (2011) Complete genome sequences of *Krokinobacter* sp. strain 4H-3-7-5 and *Lacinutrix* sp. strain 5H-3-7-4, polysaccharide-degrading members of the family Flavobacteriaceae. *J Bacteriol* **193**: 4545–4546. doi:10.1128/JB.05518-11.
- Könneke, M., Bernhard, A.E., de la Torre, J.R., Walker, C.B., Waterbury, J.B., and Stahl, D.A. (2005) Isolation of an autotrophic ammonia-oxidizing marine archaeon. *Nature* **437**: 543–546. doi:10.1038/nature03911.

- Koschorreck, M. (2008) Microbial sulphate reduction at a low pH. *FEMS Microbiol Ecol* **64**: 329–342. doi:10.1111/j.1574-6941.2008.00482.x.
- Krause, E., Wichels, A., Giménez, L., Lunau, M., Schilhabel, M.B., and Gerdts, G. (2012) Small Changes in pH Have Direct Effects on Marine Bacterial Community Composition: A Microcosm Approach. *PLoS ONE* **7**: e47035. doi:10.1371/journal.pone.0047035.
- Krulwich, T.A., Sachs, G., and Padan, E. (2011) Molecular aspects of bacterial pH sensing and homeostasis. *Nat Rev Microbiol* **9**: 330–343.
- Lehtovirta-Morley, L.E., Stoecker, K., Vilcinskas, A., Prosser, J.I., and Nicol, G.W. (2011) Cultivation of an obligate acidophilic ammonia oxidizer from a nitrifying acid soil. *Proc Natl Acad Sci USA* **108**: 15892–15897. doi:10.1073/pnas.1107196108.
- Madigan, M.T., Martinko, J.M., Parker, J., and Brock, T.D. (1997) *Biology of microorganisms*. Upper Saddle River, NJ, USA: Prentice Hall.
- Makarenkov, V., and Legendre, P. (2002) Nonlinear redundancy analysis and canonical correspondence analysis based on polynomial regression. *Ecology* **83**: 1146–1161.
- Marteinson, V.T., Bjornsdottir, S.H., Bienvenu, N., Kristjansson, J.K., and Birrien, J.L. (2010) *Rhodothermus profundus* sp. nov., a thermophilic bacterium isolated from a deep-sea hydrothermal vent in the Pacific Ocean. *Int J Syst Evol Microbiol* **60**: 2729–2734. doi:10.1099/ijs.0.012724-0.
- Martens, T., Heidorn, T., Pukall, R., Simon, M., Tindall, B.J., and Brinkhoff, T. (2006) Reclassification of *Roseobacter gallaeciensis* Ruiz-Ponte et al. 1998 as *Phaeobacter gallaeciensis* gen. nov., comb. nov., description of *Phaeobacter inhibens* sp. nov., reclassification of *Ruegeria algicola* (Lafay et al. 1995) Uchino et al. 1999 as *Marinovum algicola* gen. nov., comb. nov., and emended descriptions of the genera *Roseobacter*, *Ruegeria* and *Leisingera*. *Int J Syst Evol Microbiol* **56**: 1293–1304. doi:10.1099/ijs.0.63724-0.
- Martens-Habbena, W., Berube, P.M., Urakawa, H., de la Torre, J.R., and Stahl, D.A. (2009) Ammonia oxidation kinetics determine niche separation of nitrifying Archaea and Bacteria. *Nature* **461**: 976–979. doi:10.1038/nature08465.
- Meron, D., Atlas, E., Iasur Kruh, L., Elifantz, H., Minz, D., Fine, M., and Banin, E. (2010) The impact of reduced pH on the microbial community of the coral *Acropora eurystoma*. *ISME J* **5**: 51–60. doi:10.1038/ismej.2010.102.
- Morrow, K.M., Bourne, D.G., Humphrey, C., Botté, E.S., Laffy, P., Zaneveld, J., et al. (2014) Natural volcanic CO<sub>2</sub> seeps reveal future trajectories for host–microbial associations in corals and sponges. *ISME J*: Advanced online publication. doi:10.1038/ismej.2014.188.
- Mouchka, M.E., Hewson, I., and Harvell, C.D. (2010) Coral-associated bacterial assemblages: current Knowledge and the potential for climate-driven impacts. *Integr Comp Biol* **50**: 662–674. doi:10.1093/icb/icq061.
- Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B., et al. (2013) *vegan: community ecology package*. URL <http://CRAN.R-project.org/package=vegan>.
- Pandolfi, J.M., Connolly, S.R., Marshall, D.J., and Cohen, A.L. (2011) Projecting coral reef futures under global warming and ocean acidification. *Science* **333**: 418–422. doi:10.1126/science.1204794.
- Park, S.C., Choe, H.N., Baik, K.S., and Seong, C.N. (2012) *Aquimarina mytili* sp. nov., isolated from the gut microflora of a mussel, *Mytilus coruscus*, and emended description of *Aquimarina macrocephali*. *Int J Syst Evol Microbiol* **62**: 1974–1979. doi:10.1099/ijs.0.032904-0.
- Piontek, J., Lunau, M., Händel, N., Borchard, C., Wurst, M., and Engel, A. (2009) Acidification increases microbial polysaccharide degradation in the ocean. *Biogeosci Discuss* **6**: 11377–11400.
- Piontek, J., Borchard, C., Sperling, M., Schulz, K.G., Riebesell, U., and Engel, A. (2013) Response of bacterioplankton activity in an Arctic fjord system to elevated pCO<sub>2</sub>: results from a mesocosm perturbation study. *Biogeosciences* **10**: 297–314. doi:10.5194/bg-10-297-2013.
- R Core Team (2013) *R: A language and environment for statistical computing*. R Foundation for Statistical Computing. Vienna: Austria. R Foundation for Statistical Computing. URL: <http://www.R-project.org/>.
- Ramette, A. (2007) Multivariate analyses in microbial ecology. *FEMS Microbiol Ecol* **62**: 142–160. doi:10.1111/j.1574-6941.2007.00375.x.
- Ramette, A. (2009) Quantitative community fingerprinting methods for estimating the abundance of operational taxonomic units in natural microbial communities. *Appl Environ Microbiol* **75**: 2495–2505. doi:10.1128/AEM.02409-08.
- Raven, J., Caldeira, K., Elderfield, H., Hoegh-Guldberg, O., Liss, P., Riebesell, U., et al. (2005) *Ocean Acidification Due to Increasing Atmospheric Carbon Dioxide*. London, UK: The Royal Society. Policy document 12/05.
- Russell, B.D., Connell, S.D., Uthicke, S., Muehlehner, N., Fabricius, K.E., and Hall-Spencer, J.M. (2013) Future seagrass beds: can increased productivity lead to increased carbon storage? *Mar Pollut Bull* **73**: 463–469. doi:10.1016/j.marpolbul.2013.01.031.
- Saw, J.H., Yuryev, A., Kanbe, M., Hou, S., Young, A.G., Aizawa, S., and Alam, M. (2012) Complete genome sequencing and analysis of *Saprospira grandis* str. Lewin, a predatory marine bacterium. *Stand Genomic Sci* **6**: 84–93. doi:10.4056/signs.2445005.
- Schloss, P.D., Westcott, S.L., Ryabin, T., Hall, J.R., Hartmann, M., Hollister, E.B., et al. (2009) Introducing mothur: open-source, platform-independent, community-supported software for describing and comparing microbial communities. *Appl Environ Microbiol* **75**: 7537–7541. doi:10.1128/AEM.01541-09.
- Simpson, E.H. (1949) Measurement of diversity. *Nature* **163**: 688.
- Stahl, D.A., and de la Torre, J.R. (2012) Physiology and diversity of ammonia-oxidizing archaea. *Annu Rev Microbiol* **66**: 83–101. doi:10.1146/annurev-micro-092611-150128.
- Stocker, T.F., Dahe, Q., and Plattner, G.-K. (2013) Climate Change 2013: The Physical Science Basis. Working Group I Contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. Summary for Policymakers (IPCC, 2013).
- Tait, K., Laverock, B., and Widdicombe, S. (2014) Response of an Arctic sediment nitrogen cycling community to

- increased CO<sub>2</sub>. *Estuaries Coast* **37**: 724–735. doi:10.1007/s12237-013-9709-x.
- Tarre, S., Beliaevski, M., Denekamp, N., Gieseke, A., de Beer, D., and Green, M. (2004) High nitrification rate at low pH in a fluidized bed reactor with chalk as the biofilm carrier. *Water Sci Technol* **49**: 99–105.
- Teira, E., Fernández, A., Álvarez-Salgado, X.A., García-Martín, E.E., Serret, P., and Sobrino, C. (2012) Response of two marine bacterial isolates to high CO<sub>2</sub> concentration. *Mar Ecol Prog Ser* **453**: 27–36.
- Thompson, R.C., Norton, T.A., and Hawkins, S.J. (2004) Physical stress and biological control regulate the producer-consumer balance in intertidal biofilms. *Ecology* **85**: 1372–1382.
- Uthicke, S., Momigliano, P., and Fabricius, K.E. (2013) High risk of extinction of benthic foraminifera in this century due to ocean acidification. *Sci Rep* **3**: 1–5. doi:10.1038/srep01769.
- Vega Thurber, R., Willner-Hall, D., Rodriguez-Mueller, B., Desnues, C., Edwards, R.A., Angly, F., *et al.* (2009) Metagenomic analysis of stressed coral holobionts. *Environ Microbiol* **11**: 2148–2163. doi:10.1111/j.1462-2920.2009.01935.x.
- Venables, W.N., and Ripley, B.D. (2002) *Modern Applied Statistics with S*, 4th edn. New York, NY, USA: Springer. ISBN 0-387-95457-0.
- Walker, C.B., de la Torre, J.R., Klotz, M.G., Urakawa, H., Pinel, N., Arp, D.J., *et al.* (2010) *Nitrosopumilus maritimus* genome reveals unique mechanisms for nitrification and autotrophy in globally distributed marine crenarchaea. *Proc Natl Acad Sci USA* **107**: 8818–8823. doi:10.1073/pnas.0913533107.
- Webster, N.S., Smith, L.D., Heyward, A.J., Watts, J.E., Webb, R.I., Blackall, L.L., and Negri, A.P. (2004) Metamorphosis of a Scleractinian coral in response to microbial biofilms. *Appl Environ Microbiol* **70**: 1213–1221. doi:10.1128/AEM.70.2.1213-1221.2004.
- Witt, V., Wild, C., Anthony, K.R., Diaz-Pulido, G., and Uthicke, S. (2011) Effects of ocean acidification on microbial community composition of, and oxygen fluxes through, biofilms from the Great Barrier Reef. *Environ Microbiol* **13**: 2976–2989. doi:10.1111/j.1462-2920.2011.02571.x.
- Yoon, J.H., Kang, S.J., Lee, S.Y., and Oh, T.K. (2007) *Phaeobacter daeponensis* sp. nov., isolated from a tidal flat of the Yellow Sea in Korea. *Int J Syst Evol Microbiol* **57**: 856–861. doi:10.1099/ijs.0.64779-0.
- Zak, D.R., Grigal, D.F., and Ohmann, L.F. (1993) Kinetics of microbial respiration and nitrogen mineralization in Great Lakes forests. *Soil Sci Soc Am J* **57**: 1100–1106.
- Fig. S1.** Map of sampling area (map from Google 2014), with the seeping sites Upa-Upasina (A), Esa'Ala (B) and Dobu (C). The coordinates for each sample can be found in Table S1.
- Fig. S2.** Numbers of bacterial ARISA OTU per seawater pH category, i.e. control (pH 8.3–8.1), medium (pH 8.1–7.7) and high (pH 7.7–6.9) pCO<sub>2</sub> sites.
- Fig. S3.** Non-metric multidimensional scaling ordination of Bray–Curtis dissimilarity matrices based on bacterial community data (MPTS OTU<sub>3%</sub>) from samples of 2010 and 2011 (dotted line). The size of the symbol dots is a (non-linear) representation of differences in pCO<sub>2</sub> exposure at the respective sites. Coloured shapes highlight the three CO<sub>2</sub> impact groups 'control' (grey), 'medium pCO<sub>2</sub>' (blue) and 'high pCO<sub>2</sub>' (red).
- Fig. S4.** Dominant bacterial phyla (A) and classes (B) at the studied sites. The taxonomic assignments are based on MPTS sequences using the SILVA 16S rRNA reference database. At the top the site names with the respective seawater pH are indicated.
- Fig. S5.** Dominant archaeal phyla (A) and classes (B) at the studied sites. The taxonomic assignment is based on MPTS sequences using the SILVA 16S rRNA reference database. At the top the site names with the respective seawater pH are indicated.
- Table S1.** Overview of all 24 samples with time of sampling, sampling location, sampling depth, type of molecular analysis and measurements of environmental parameters.
- Table S2.** Summary of bacterial and archaeal sample diversity, based on ARISA and MPTS and abundance of rare bacterial and archaeal MPTS OTUs.
- Table S3.** Number of bacterial and archaeal OTUs and taxa, characterized by MPTS sequences which either uniquely occur at samples from the categories 'high', 'medium' and 'control pCO<sub>2</sub>' or show linear or quadratic relationships with pH.
- Table S4.** Bacterial taxa and MPTS OTUs, which were found to significantly increase in terms of relative sequence abundances with increasing seawater pCO<sub>2</sub>.
- Table S5.** Archaeal taxa and MPTS OTUs, which were found to significantly increase in terms of relative sequence abundances with increasing seawater pCO<sub>2</sub>.
- Table S6.** Bacterial taxa and MPTS OTUs, which were found to have a significant quadratic relationship with increasing pCO<sub>2</sub>.
- Table S7.** Archaeal taxa and MPTS OTUs, which were found to have a significant quadratic relationship with increasing pCO<sub>2</sub>.
- Text S1.** Taxa with no obvious ecological functions, Quadratic relationships of microbial MPTS sequence occurrence with pH.

## Supporting information

Additional Supporting Information may be found in the online version of this article at the publisher's web-site: