

# Trends in tuna carbon isotopes suggest global changes in pelagic phytoplankton communities

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## Abstract

Considerable uncertainty remains over how increasing atmospheric CO<sub>2</sub> and anthropogenic climate changes are affecting open-ocean marine ecosystems from phytoplankton to top predators. Biological time series data are thus urgently needed for the world's oceans. Here, we use the carbon stable isotope composition of tuna to provide a first insight into the existence of global trends in complex ecosystem dynamics and changes in the oceanic carbon cycle. From 2000 to 2015, considerable declines in δ<sup>13</sup>C values of 0.8‰–2.5‰ were observed across three tuna species sampled globally, with more substantial changes in the Pacific Ocean compared to the Atlantic and Indian Oceans. Tuna recorded not only the Suess effect, that is, fossil fuel-derived and isotopically light carbon being incorporated into marine ecosystems,

but also recorded profound changes at the base of marine food webs. We suggest a global shift in phytoplankton community structure, for example, a reduction in  $^{13}\text{C}$ -rich phytoplankton such as diatoms, and/or a change in phytoplankton physiology during this period, although this does not rule out other concomitant changes at higher levels in the food webs. Our study establishes tuna  $\delta^{13}\text{C}$  values as a candidate essential ocean variable to assess complex ecosystem responses to climate change at regional to global scales and over decadal timescales. Finally, this time series will be invaluable in calibrating and validating global earth system models to project changes in marine biota.

#### KEYWORDS

albacore tuna, Atlantic Ocean, bigeye tuna, biogeochemical cycles, carbon cycle, Indian Ocean, Pacific Ocean, phytoplankton, Suess effect, yellowfin tuna

## 1 | INTRODUCTION

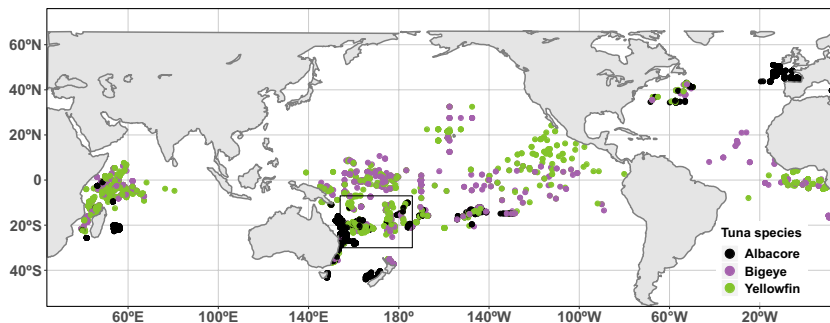
Over the past 50 years, 90% of the heat associated with global warming, and 30% of the fossil fuel carbon emissions have been absorbed by the oceans (Le Quere et al., 2018). Such processes are predicted to severely impact marine biota (Poloczanska et al., 2016) through enhanced ocean stratification and acidification. Unfortunately, there are large uncertainties on how oceanic ecosystems have changed or may change in the future. For example, the current generation of earth system models simulates a wide range of future changes in global ocean net primary productivity (NPP), with both increases and decreases of up to 20% by 2100 (Bopp et al., 2013; Kwiatkowski et al., 2017), highlighting large discrepancies in the trends of simulated NPP. Only a limited number of empirical datasets record trends in the phytoplankton community composition or physiology (Gregg & Rousseaux, 2014; Gregg, Rousseaux, & Franz, 2017; Rousseaux & Gregg, 2015). The magnification of relative changes in phytoplankton dynamics across trophic levels has rarely been investigated with few available empirical methods capable of quantifying ecosystem-level responses. Biological time series datasets are imperative for understanding past responses of the world's oceans and for quantifying uncertainty in future climate projections (Bonan & Doney, 2018).

Carbon stable isotopes ( $\delta^{13}\text{C}$  values or  $^{13}\text{C}/^{12}\text{C}$ ) have been used to reconstruct the oceanic carbon cycle using direct measurements or marine archives (e.g., marine sediments, corals) from paleoclimates to the current anthropogenic perturbation (Ehleringer, Buchmann, & Flanagan, 2000; Freeman & Hayes, 1992; Keeling, 2017; Wu et al., 2018). Since the Industrial Revolution, the rise in atmospheric  $\text{CO}_2$  has been accompanied by a decrease in the carbon isotope ratio of atmospheric  $\text{CO}_2$ , known as the Suess effect (Keeling, 1979). This decrease is attributed to the atmospheric release of isotopically light carbon from fossil fuel combustion. Due to the oceanic uptake of this  $^{13}\text{C}$ -depleted  $\text{CO}_2$ , the oceanic  $\delta^{13}\text{C}$  value of dissolved inorganic carbon ( $\delta^{13}\text{C}_{\text{DIC}}$ ) is decreasing (Quay, Sonnerup, Munro, & Sweeney, 2016; Quay et al., 2007). Changes in  $\delta^{13}\text{C}_{\text{DIC}}$  values are recorded in phytoplankton  $\delta^{13}\text{C}$  values after

accounting for an isotopic fractionation factor associated with photosynthesis (defined as  $\epsilon_p$ ). Isotopic fractionation is dependent on seawater characteristics, phytoplankton composition, and physiology. The primary factors that are believed to affect the isotopic values of phytoplankton are: (a) the concentration and  $\delta^{13}\text{C}$  values of dissolved  $\text{CO}_2$  ( $[\text{CO}_2]_{\text{aq}}$ ; Fry, 1996; Laws, Popp, Bidigare, Kennicutt, & Macko, 1995; Popp et al., 1998); (b) phytoplankton community composition and cell morphology (Popp et al., 1998); and (c) cellular growth rate (Bidigare et al., 1997; Fry, 1996). Secondary physiological traits (e.g., decreases in bicarbonate uptake or in carbon-concentrating mechanism activity) can also impact isotopic values, but are difficult to model (Cassar, Laws, & Popp, 2006).

Carbon isotopic changes at the base of food webs are transferred to higher trophic levels with values increasing slightly (typically 0.5‰–1‰) with each trophic transfer (Fry, 2006; Graham et al., 2010). Metabolically active tissues of consumers (e.g., fish muscle) integrate the stable isotope values of this base through their diet (Cherel & Hobson, 2007; Graham et al., 2010). While nitrogen isotope ( $\delta^{15}\text{N}$ ) values are commonly used to investigate changes in trophic levels,  $\delta^{13}\text{C}$  values provide information on animal diets and on spatial variations at the base of food webs (Cherel & Hobson, 2007; MacKenzie, Longmore, Preece, Lucas, & Trueman, 2014; Trueman, MacKenzie, & Palmer, 2012). Historical studies focusing on baseline changes have examined accretionary bioarchives that suffer little degradation after formation such as keratin baleen plates, feathers or teeth dentin of marine consumers that reflect the food they ingest and therefore, the  $\delta^{13}\text{C}$  values of phytoplankton (Jaeger & Cherel, 2011; Newsome et al., 2007; Schell, 2001). These studies demonstrate the utility of isotope measurements to reconstruct past and present ocean primary productivity, and provide evidence of past climate changes at regional scales (Hobson, Sinclair, York, Thomason, & Merrick, 2004; Newsome et al., 2007; Schell, 2001). Finally, metabolically inert but inorganic accretionary structures (e.g., bivalve shells, coral skeletons, sclerosponges, or fish otoliths) can also reflect the  $\delta^{13}\text{C}_{\text{DIC}}$  value of the environment across their lifetime (Fraile et al.,

**FIGURE 1** Map of global study area and locations of 4,477 samples for three tuna species. The black square delineates the New Caledonia–Fiji region used for a focused spatial and temporal analysis



2016; Swart, 2010) as they usually precipitate in equilibrium with seawater, although vital effects can complicate environmental reconstruction (Lorrain, 2004; McConnaughey & Gillikin, 2008). Similarly,  $\delta^{13}\text{C}$  values of metabolically active tissues may reflect trends in physio-chemical processes ( $\text{CO}_2$  and  $\delta^{13}\text{C}_{\text{aq}}$  values) and biological processes (phytoplankton  $\delta^{13}\text{C}$  values).

The aim of our study was to assess trends in a time series of stable isotope values of metabolically active tuna tissues, and to test if this could be used to detect ecosystem-level responses over decadal time scales at regional to global scales. For this purpose, we analyzed  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values of muscle samples from three species of tuna (yellowfin tuna, *Thunnus albacares*; bigeye tuna, *T. obesus*; and albacore tuna, *T. alalunga*) collected throughout tropical, subtropical, and temperate oceans from 2000 to 2015 ( $n = 4,477$ ; Figure 1). Each of these species has different vertical and foraging distributions (from surface to mesopelagic depths) (Olson et al., 2016). Therefore, our study seeks to resolve broad horizontal and vertical spatial patterns in oceanic food webs. As tuna are widely distributed and harvested globally (Majkowski, 2007), they are good candidates to study how observed and suspected changes in physical and biological processes at global and ocean basin scales may be reflected in consumer  $\delta^{13}\text{C}$  values. We developed a theoretical model to decompose the observed temporal changes in consumer  $\delta^{13}\text{C}$  values into putative causal contributors. The model accounted for (a) known temporal trends in fossil fuel-derived carbon (the Suess effect) and  $\text{CO}_2$  availability; (b) possible changes in phytoplankton dynamics including community composition and growth rates; and (c) potential changes in the trophic fractionation factor. Our study, which focuses on carbon but draws on nitrogen isotopes to assess potential changes in tuna trophic positions, suggests large-scale shifts in phytoplankton communities from 2000 to 2015.

## 2 | MATERIALS AND METHODS

### 2.1 | Tuna carbon isotope data

We assembled a global database using published and unpublished regional carbon isotope studies resulting in 4,477 records from 2000 to 2015. Details on isotopic methods and predator sampling

are provided in Pethybridge et al. (2018) who analyzed the same global dataset but for  $\delta^{15}\text{N}$  values. As for  $\delta^{15}\text{N}$  and other global compilation studies (Bird et al., 2018), we assumed that the agreement between  $\delta^{13}\text{C}$  values generated across different laboratories was  $<0.2\text{‰}–0.3\text{‰}$ . Tuna size (fork length, in cm) was measured for each individual. Tuna were sampled from three ocean basins (Atlantic, Indian, and Pacific Oceans) with albacore tuna occupying more temperate waters compared to the tropical yellowfin and bigeye tunas (Figure 1; Olson et al., 2016).

The Pacific Ocean had the most extensive sampling with 2,504 individuals and no gap from 2000 to 2015, except for albacore where data were not available for 4 years. In the Indian and Atlantic Oceans, data were more scattered (Table S1). Pelagic tuna are mobile predators and the stable carbon isotopic composition of their muscle tissue represents an integration of their foraging environment over approximately 6 months to 1 year (Houssard et al., 2017). Tuna muscle tissue  $\delta^{13}\text{C}$  values were corrected for lipids in all samples either with chemical extraction or using a mass balance equation for elevated lipid content samples ( $\text{C}/\text{N} > 3.5$ ) with parameters derived from Atlantic bluefin tuna (*T. thynnus*) muscle (Logan et al., 2008).

### 2.2 | Temporal trends in tuna $\delta^{13}\text{C}$ values

Time series analyses based on multiple linear regression analysis, performed using the R-3.2.4 software (R Development Core Team, 2016) and the nlme package (Pinheiro et al., 2018), were used to examine and test for significant linear trends in tuna carbon isotope values. To ensure that tuna length (size) did not have any effect on potential temporal trends, an interaction between size and year was tested and was not found to be significant for the Pacific or Atlantic Ocean. For the global dataset, we tested a model with tuna size included, by species and ocean basins, and then fitted a model explaining the residuals of this first model as a function of year. The slopes were similar to those obtained without the effect of size included, meaning that the addition of size does not change the observed patterns and that this factor has a small impact on temporal trends in  $\delta^{13}\text{C}$  values. We finally tested for three variables: year (quantitative), ocean with three levels (Atlantic, Indian, and Pacific Oceans), and tuna species with three levels (albacore, bigeye, and yellowfin). All combinations were tested and the final model was chosen using the Akaike information criterion. We added an autocorrelation structure:

a one-degree autoregressive integrated moving average (ARIMA; Pinheiro et al., 2018) via the gls function fitted by groups of tuna sampled at the same date and position. Autocorrelation structures on residuals were checked with an autocorrelation function. Finally, to account for possible spatial biases (a) in years of sampling according to locations (Figure S1) or (b) due to baseline isotopic variations across space (McMahon, Ling Hamady, & Thorrold, 2013), tuna  $\delta^{13}\text{C}$  trends were also determined at a smaller spatial scale by considering one region where sufficient data were available per year, that is, New Caledonia and Fiji (see Figure 1 for selected region area) and with similar isotope values at the base of the food web (Houssard et al., 2017; Magozzi, Yool, Vander Zanden, Wunder, & Trueman, 2017). Furthermore, all New Caledonia and Fiji samples were analyzed in the same laboratory and time period.

### 2.3 | Modeling the factors influencing tuna $\delta^{13}\text{C}$ values

We developed a theoretical model to explain the potential effects of various factors and processes known to explain trends in tuna  $\delta^{13}\text{C}$  values. First, we considered the isotope value of phytoplankton ( $\delta^{13}\text{C}_p$ ) that has been shown to be driven by the magnitude of carbon isotopic fractionation during photosynthesis ( $\epsilon_p$ ) and the isotope value of  $\text{CO}_2$  ( $\delta^{13}\text{C}_{\text{aq}}$ , that is, the Suess effect), with  $\epsilon_p$  dependent on the carbon isotope fractionation associated with carbon fixation ( $\epsilon_f$ ) and the specific growth rate ( $\mu$ ; Laws et al., 1995).

$$\delta^{13}\text{C}_p = \frac{1}{\left(1 + \frac{\epsilon_p}{1,000}\right)} \left(\delta^{13}\text{C}_{\text{aq}} - \epsilon_p\right), \quad (1)$$

with

$$\epsilon_p = \epsilon_f - \frac{b\mu}{\text{CO}_2}, \quad (2)$$

where  $b$  is a constant (mM/day) reflecting the degree of dependence of fractionation on the  $\text{CO}_2$  concentration, and is believed to vary between species and as a function of growth conditions (Bidigare et al., 1997; Cullen, Rosenthal, & Falkowski, 2001). While the parameter values are arbitrary, they are within the range of values reported in the literature (Table 1). An initial value of 120 was used for  $b$  which is consistent with the range of values (52.6–137.9) from Popp et al. (1998) for *Emiliana huxleyi* ( $\epsilon_p = 24.6\text{--}137.9 \mu/\text{CO}_2$ ) the most common species of coccolithophore globally (Beardall & Raven, 2013), and *Phaeodactylum tricornutum* ( $\epsilon_p = 25.5\text{--}52.6 \mu/\text{CO}_2$ ), a diatom mostly used in laboratory studies but not representative of the global ocean. For the intrinsic fractionation during photosynthesis by the enzyme Rubisco ( $\epsilon_f$ ), we used a value of 25‰, which has been estimated to range between ~22‰ and 30‰ depending on species (e.g., Popp et al., 1998), with values as low as 11‰ for the Rubisco of the coccolithophore *E. huxleyi* (Boller, Thomas, Cavanaugh, & Scott, 2011). The value of 25‰ for  $\epsilon_f$  has often been proposed (Bidigare et al., 1997) in studies trying to understand temporal trends in marine chronologies (Schell, 2001).

Factors ( $x$ )	Starting value or equation used	Sensitivity $\left(\frac{d\delta^{13}\text{C}_{\text{tuna}}}{\delta^{13}\text{C}_{\text{tuna}} / \frac{dx}{x}}\right)$	Imposed change %	Tuna $\Delta\delta^{13}\text{C}$ , % 16-years	% Change explained			
					NC-Fiji	PO	AO	IO
$\delta^{13}\text{C}_{\text{aq}}$	Quay et al. (2016)	-0.08	NA	0.30	14	12	22	38
$\text{CO}_2$	Cassar et al. (2006)	0.21	NA	0.06	3	2	4	8
	Bidigare (1997)		NA	0.27	13	11	20	35
Growth rate ( $\mu$ )	0.3	-0.2	-5	0.20	10	8	14	26
			-10	0.30	14	12	22	38
			-15	0.48	23	19	35	62
Carbon fixation fractionation factor ( $\epsilon_f$ )	25	1.53	+1	0.24	12	10	17	31
			+2	0.48	23	19	35	61
			+3	0.72	35	29	52	92
			+5	1.20	58	48	87	154
$b$ Factor	120	-0.2	-2	0.10	5	4	7	13
			-4	0.20	10	8	14	26
			-8	0.40	19	16	29	51
			-10	0.50	24	20	36	64
Trophic fractionation factor ( $\epsilon_{f_c}$ )	4	-0.23	-5	0.20	10	8	14	26
			-10	0.40	19	16	29	51
			-15	0.59	29	24	43	76

**TABLE 1** Parameter and time series data used to run various scenarios of imposed changes in phytoplankton dynamics, and the effects on the different spatial areas examined (New Caledonia–Fiji [NC-Fiji] region, Pacific Ocean [PO], Atlantic Ocean [AO], and Indian Ocean [IO]). For example,  $\delta^{13}\text{C}_{\text{aq}}$  explained 12% of the tuna  $\delta^{13}\text{C}$  decrease  $\Delta\delta^{13}\text{C}$ , in 16 years in the PO, while an imposed 5% change in carbon fixation fractionation factor ( $\epsilon_f$ ) explained 48% of  $\Delta\delta^{13}\text{C}$ , this factor having the largest sensitivity (1.53) compared to all factors tested (see Section 2 for more details)

While precise estimates of  $\epsilon_f$  or  $b$  are not available, this parametrization provides a quantitative demonstration of how even small changes in phytoplankton community composition or physiology may influence tuna muscle  $\delta^{13}\text{C}$  values, and hence emergent signals of change (Table 1). Growth rate  $\mu$  was set at  $0.3 \text{ day}^{-1}$  as it is the median value at Station ALOHA (Laws et al., 2013) in the central North Pacific (Hawaii Ocean Time-series) and was also used by several authors (Cullen et al., 2001). Ranges for  $\mu$  from  $0.1$  to  $1 \text{ day}^{-1}$  have been reported in the literature (Boyd et al., 2013; Laws et al., 2013). We proposed a decrease in growth rate of up to 15% over the 2000–2015 study period (from 0.30 to 0.26), which is on the high end of observed modern changes (Gregg et al., 2017) and predicted decreases for the future (Kwiatkowski et al., 2017). The  $\delta^{13}\text{C}_{\text{aq}}$  values were taken from Station ALOHA (Quay et al., 2016).

Changes in tuna  $\delta^{13}\text{C}$  values can in turn be described by the following equation:

$$\delta^{13}\text{C}_{\text{tuna}} = \frac{1}{\left(1 + \frac{\epsilon_{\text{fc}}}{1,000}\right)} \left( \delta^{13}\text{C}_{\text{p}} - \epsilon_{\text{fc}} \right), \quad (3)$$

where  $\epsilon_{\text{fc}}$  is the overall fractionation associated with trophic transfers and is considered to be low ( $\sim 0.5\%$  to  $1.8\%$  per trophic level), therefore we used  $4\%$  for tuna that are considered to be at a trophic position of  $\sim 4$  (Olson et al., 2016). As a comparison, Bird et al. (2018) found an average difference of  $4.6\%$  between phytoplankton and sharks on a global scale.

Combining Equation (2) with Equation (3) leads to:

$$\delta^{13}\text{C}_{\text{tuna}} = \frac{1}{\left(1 + \frac{\epsilon_{\text{fc}}}{1,000}\right)} \left( \left( \frac{1}{\left(1 + \frac{\left(\epsilon_f - \frac{b\mu}{\text{CO}_2}\right)}{1,000}\right)} \right) \left( \delta^{13}\text{C}_{\text{aq}} - \left( \epsilon_f - \frac{b\mu}{\text{CO}_2} \right) \right) \right) - \epsilon_{\text{fc}}. \quad (4)$$

We used two different parametrizations developed in the literature on the effect of the concentration of  $\text{CO}_{2\text{aq}}$  on the isotope values of phytoplankton (Bidigare et al., 1997; Cassar et al., 2006). Indeed, while the first parametrization (Bidigare et al., 1997; Laws et al., 1995) provides quantitative intuition for the dependence of tuna  $\delta^{13}\text{C}$  values on phytoplankton physiology, it does not account for additional factors influencing the phytoplankton isotope values, including changes in the carbon source (bicarbonate vs.  $\text{CO}_2$ ), deactivation of carbon-concentrating mechanisms in response to increased  $\text{CO}_2$  availability, and changes in the growth conditions (nutrient vs. light limitation). For completeness, we also show the predictions based on the parametrization presented in Cassar et al. (2006). The sensitivity of tuna  $\delta^{13}\text{C}$  values to each factor was assessed by calculating the ratio of the percentage change of tuna  $\delta^{13}\text{C}$  values to percentage change of each factor. The slope ('m') of each curve is related to the ratio of the percentage change of tuna  $\delta^{13}\text{C}$  values to percentage change of each factor according to the following equation:

$$\left( \frac{d\delta^{13}\text{C}_{\text{tuna}} / dx}{\delta^{13}\text{C}_{\text{tuna}} / x} \right) = m \frac{x}{\delta^{13}\text{C}_{\text{tuna}}}. \quad (5)$$

This sensitivity analysis examines the influence of one parameter at a time on tuna isotopic composition with assumed initial values for each parameter based on literature values. To reinforce this analysis, we conducted a Bayesian approach that takes into account the uncertainty of all parameters simultaneously to explain tuna isotopic composition with ranges and uncertainties taken from literature values (Analysis S1). To reconstruct the observed trends in tuna  $\delta^{13}\text{C}$  values, a number of scenarios were run to simulate percentage changes in the phytoplankton parameters  $\mu$ ,  $b$ , and  $\epsilon_f$  or  $\epsilon_{\text{fc}}$  (Table 1). These scenarios were used to resolve competing hypotheses for the observed patterns.

## 3 | RESULTS

### 3.1 | Trends in $\delta^{13}\text{C}$ values

Over the entire record (Figure 1), individual tuna  $\delta^{13}\text{C}$  values ranged from  $-19.9\%$  to  $-12.9\%$ . Mean annual  $\delta^{13}\text{C}$  values decreased by  $0.8\%$  to  $2.5\%$  within species and ocean basins from 2000 to 2015 (Table 2; Figure 2). These negative trends were significant, with similar observed slopes for each tuna species by ocean basin ( $p < .0001$ ; Table 2). The largest decrease was observed for the Pacific Ocean and the lowest in the Indian Ocean, with the decrease in the Atlantic Ocean being intermediate. For reference, we showed a global decreasing trend acknowledging that most of our observations were from the Pacific (56%) (Table 2; Figure 2;  $1.8\%$  decrease from 2000 to 2015), while observations from the Indian and Atlantic Oceans each comprised 22% of the data (Table 2).

In the region of New Caledonia and Fiji, where we have the most complete record to derive a temporal trend (Figure S1), the same decreasing pattern in tuna  $\delta^{13}\text{C}$  values was observed for all three species ( $2.1\%$  decrease between 2000 and 2015; Figure S2) as over the broader Pacific region ( $2.5\%$  decrease). No temporal changes in  $\delta^{15}\text{N}$  values were observed for the three tuna species in the New Caledonia and Fiji regions (Figure S3), suggesting no significant tuna trophic position changes over the record. Some weak trends in  $\delta^{15}\text{N}$  values were found for some species and ocean basins, which could arise from the interaction with other confounding factors such as tuna size and location (Figure S4). Two likely explanations for the observed tuna  $\delta^{13}\text{C}$  trends are discussed below.

### 3.2 | Accounting for the observed Suess effect and $\text{CO}_2$ availability

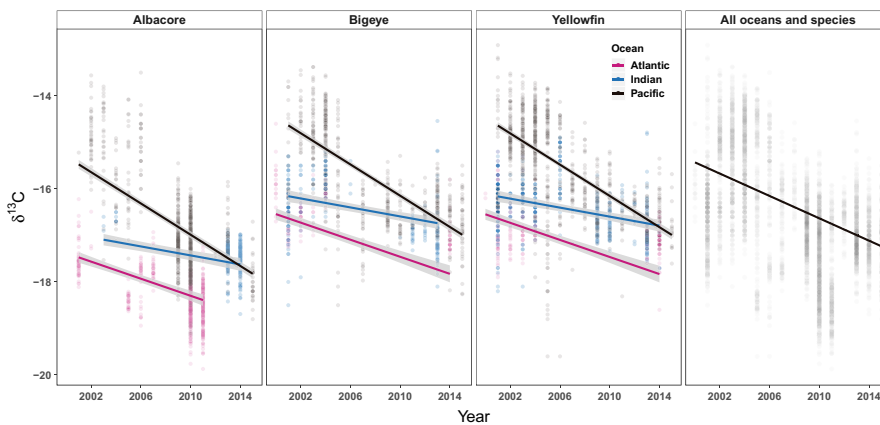
Reported declines in  $\delta^{13}\text{C}_{\text{aq}}$  values at Station ALOHA during the 2000–2015 period ( $-0.3\%$ ) explained 14% of the decrease in tuna  $\Delta\delta^{13}\text{C}$  values observed in our New Caledonia–Fiji region (Figure 3a).

Assuming similar Suess effects in the other ocean basins, 12%, 22%, and 38% of the decrease in tuna  $\delta^{13}\text{C}$  values in the Pacific, Atlantic, and Indian Oceans, respectively, can be explained by  $\delta^{13}\text{C}_{\text{aq}}$  (Table 1).

An increase in the concentration of  $\text{CO}_{2\text{aq}}$  observed at Station ALOHA also leads to a decrease in tuna  $\delta^{13}\text{C}$  values by increasing carbon isotopic fractionation during photosynthesis  $\epsilon_p$  (Equation 2).

Ocean basin/ region	Tuna species	Intercept (in 2000)	Slope	Slope stand- ard error	$r^2$	Temporal change (2000– 2015, in ‰)	$n$
Atlantic	Albacore	-17.4	-0.092	0.0085	69.1	-1.38	608
	Bigeye	-16.6					126
	Yellowfin	-16.6					256
Indian	Albacore	-16.8	-0.052	0.0077	69.1	-0.78	248
	Bigeye	-16.1					237
	Yellowfin	-16.1					498
Pacific	Albacore	-15.2	-0.166	0.0048	69.1	-2.49	878
	Bigeye	-14.5					645
	Yellowfin	-14.5					981
New Caledonia- Fiji	Albacore	-15.3	-0.138	0.0095	42.1	-2.07	364
	Bigeye	-14.8					120
	Yellowfin	-14.9					331
Global	All	-15.4	-0.120	0.0057	22.1	-1.80	4,477

**TABLE 2** Regression analysis output including the slope and intercept for each tuna species and ocean basin. Only one value is shown when similar for several species



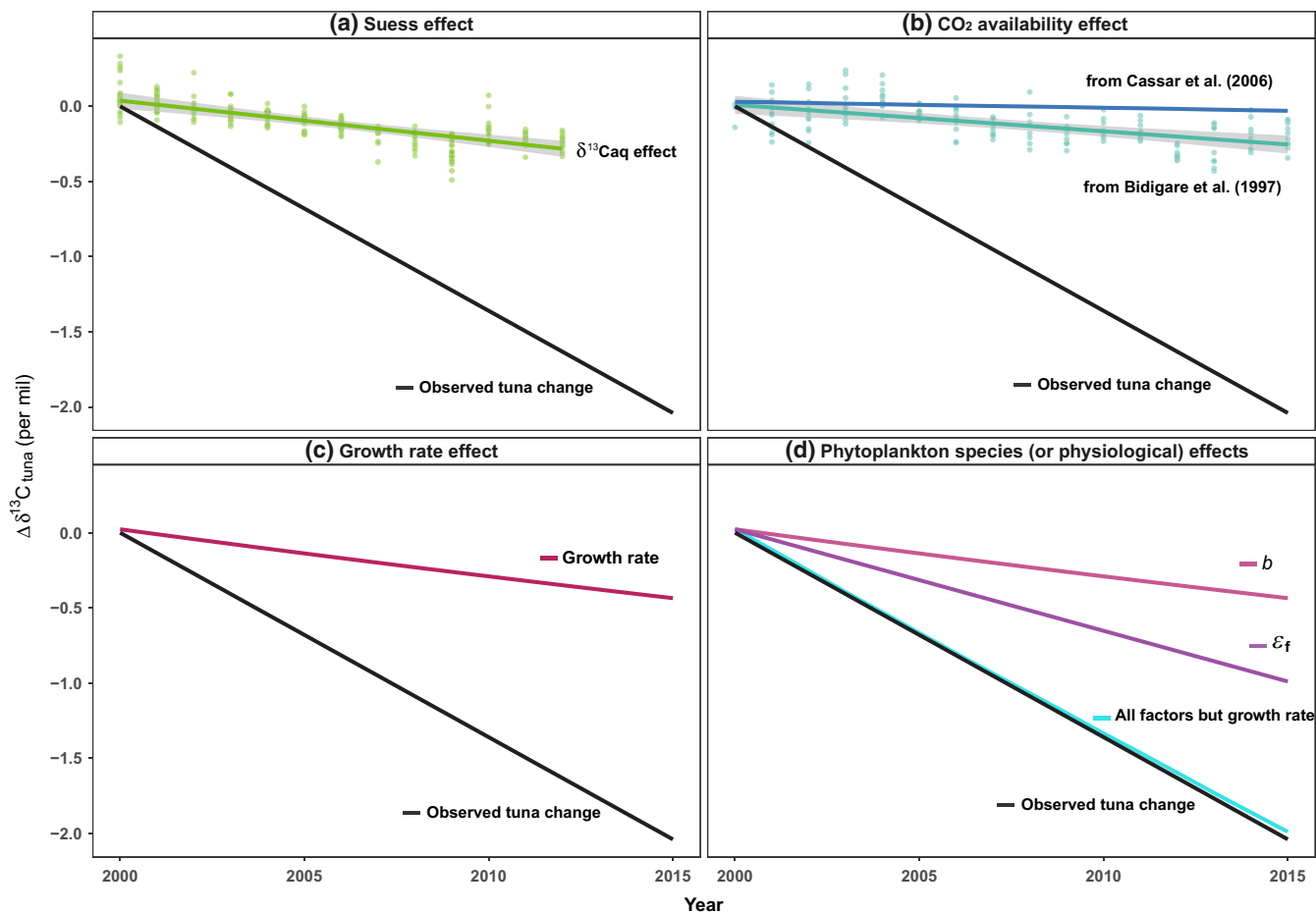
**FIGURE 2** Time series of tuna muscle tissue  $\delta^{13}\text{C}$  values (‰) with observations divided by ocean basin. The shaded area along the linear fit corresponds to a 95% confidence interval

Depending on the parametrization used to account for this effect (Bidigare et al., 1997; Cassar et al., 2006), 2%–35% of the trend in tuna  $\delta^{13}\text{C}$  values can be explained by changes in  $\text{CO}_2$  availability in the New Caledonia–Fiji region (Figure 3b). Using the larger degree of change in response to  $\text{CO}_2$  availability (Bidigare et al., 1997), percentages of change similar to those from the Suess effect can be explained across the different ocean basins. The additive impacts of  $\text{CO}_2$  availability and the Suess effect in explaining tuna  $\Delta\delta^{13}\text{C}$  values are 23%, 41%, and 73% in the Pacific, Atlantic, and Indian Oceans, respectively. If the  $\text{CO}_2$  availability effect reported by Cassar et al. (2006) is used, in addition to the Suess effect, then only 14%, 26%, and 46% of the  $\Delta\delta^{13}\text{C}$  decrease in the Pacific, Atlantic, and Indian Oceans, respectively, can be explained.

### 3.3 | Hypothesized changes in phytoplankton dynamics

According to theoretical models (see Section 2), the then unexplained temporal changes in tuna  $\delta^{13}\text{C}$  values (~27%–86%) must be

related to (a) a decrease in phytoplankton cellular growth rates ( $\mu$ ) or physiology (e.g., carbon-concentrating mechanism activity); and/or (b) potential changes in phytoplankton communities (through changes in species-dependent parameters  $b$  and  $\varepsilon_f$ ) or in the trophic fractionation factor  $\varepsilon_{fc}$ . Based both on the sensitivity analysis and the Bayesian inference, variations in the growth rate ( $\mu$ ) and in the trophic fractionation factor ( $\varepsilon_{fc}$ ) have a small effect on tuna isotope values (Table 1; Analysis S1). As an example, an imposed substantial 15% decrease in the growth rate  $\mu$  over 16 years resulted in only a 0.48‰ decrease in tuna  $\delta^{13}\text{C}$  values (Figure 3c; around 20% of the total decrease in the New Caledonia–Fiji region and in the broader Pacific Ocean). This effect is larger in the Atlantic (35%) and Indian (62%) Oceans. However, more reasonable declines of 5% and 10% of  $\mu$  over 16 years resulted in smaller decreases of 0.2‰ and 0.3‰ in tuna  $\delta^{13}\text{C}$  values from 2000 to 2015, respectively (Table 1), which explained 8%–38% of this overall signal in various ocean basins. Variations in the trophic fractionation factor  $\varepsilon_{fc}$  were of similar order (Table 1) with large decreases of this parameter needed to explain the tuna  $\delta^{13}\text{C}$  pattern.

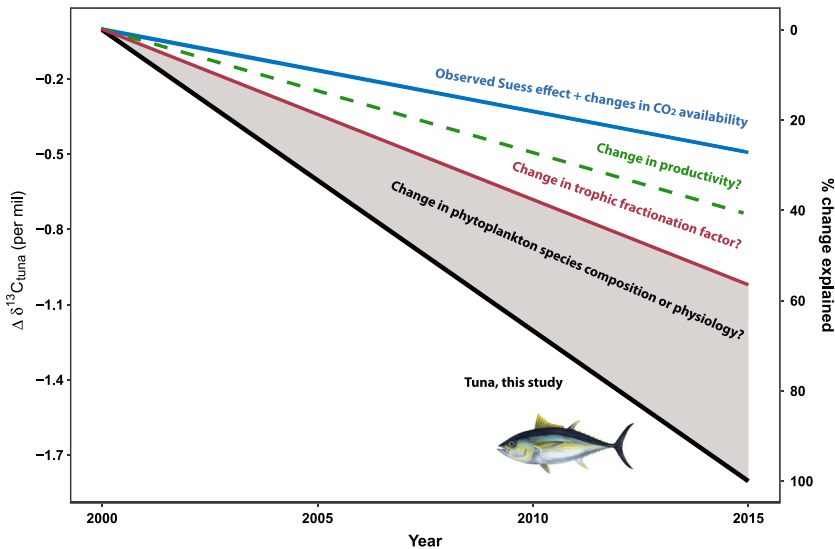


**FIGURE 3** Predicted (color line) versus observed (black line) changes in tuna muscle  $\delta^{13}\text{C}$  values ( $\Delta\delta^{13}\text{C}$ , ‰) in the New Caledonia–Fiji region of the Pacific Ocean as a function of various processes. (a) The Suess effect. (b) Increase in  $\text{CO}_{2\text{aq}}$  in seawater under two scenarios based on different parametrizations in the literature (Bidigare et al. 1997, Cassar et al., 2006; see Section 2 for details). (c) A decrease in phytoplankton cellular growth rate of 15%. (d) A change of 5% for the carbon fixation fractionation factor  $\epsilon_f$  and 10% for the constant  $b$  used to calculate carbon isotope fractionation during photosynthesis (see Section 2 for details), and also all factors considered together, except growth rate (blue line = Suess effect +  $\text{CO}_{2\text{aq}}$  from Bidigare et al., 1997 +  $b + \epsilon_f$ )

The carbon fixation fractionation factor ( $\epsilon_f$ ) and  $b$  values can vary widely among phytoplankton species (Popp et al., 1998). The sensitivity analysis and the Bayesian model (that takes into account a large range of values for these parameters) showed that the carbon fixation fractionation factor  $\epsilon_f$  had the largest effect on the tuna isotope values compared to all other factors (Table 1; Analysis S1). As an example, we arbitrarily set the changes to 5% for  $\epsilon_f$  and 10% for  $b$  (Figure 3d) to reflect their differential impact on tuna  $\Delta\delta^{13}\text{C}$  values. This small 5% increase in  $\epsilon_f$  resulted in a large decrease in tuna carbon isotope values of 1.2‰ (i.e., ~50% of the tuna  $\Delta\delta^{13}\text{C}$  in the New Caledonia–Fiji region and the Pacific Ocean) (Figure 3d; Table 1). In comparison, a 10% decline in ‘ $b$ ’ values only caused a 0.5‰ decline in tuna  $\delta^{13}\text{C}$  values, which explained 20%–36% in the Pacific and Atlantic Oceans, and 64% in the Indian Ocean.

After accounting for the Suess and  $\text{CO}_2$  availability effects, several permutations for the parameters reflecting productivity ( $\mu$ ) and species composition (with changes in  $\epsilon_f$  and  $b$  combined) may account for the remaining  $\Delta\delta^{13}\text{C}$  changes. The combination of the Suess effect, the effect of increasing  $[\text{CO}_2]_{\text{aq}}$  on  $\epsilon_p$ , and a change

of 5%–10% in species-specific parameters (10% for  $b$  and 5% for  $\epsilon_f$ ) with no change in productivity or the trophic fractionation factor  $\epsilon_{\text{fc}}$ , produced a 2.1‰ decrease in tuna  $\delta^{13}\text{C}$  values, consistent with the observed change in the Pacific Ocean (Figure 3d). In the Pacific Ocean, where we have the most robust dataset, changes in phytoplankton parameters seem to have occurred, unless we assume that growth rates have changed by >70%. If we assume that no changes in growth rates and  $\epsilon_{\text{fc}}$  have occurred and use the Bidigare et al. (1997) parametrization, then more than 60% of  $\Delta\delta^{13}\text{C}$  has to be explained by a change in species composition in the Pacific and Atlantic Oceans, against only 27% in the Indian Ocean (Table 1). The use of the Cassar et al. (2006) parametrization implies even larger changes in species composition. Averaging all tuna species and all ocean basins, and both parametrizations used to calculate carbon fractionation from phytoplankton, the global trend in tuna  $\delta^{13}\text{C}$  values ( $\Delta\delta^{13}\text{C}$ ) can, for example, be explained by (a) the observed Suess effect and increases in  $\text{CO}_{2\text{aq}}$  (up to 26%); (b) a 5% decrease in productivity (11%), a 10% decrease in the trophic fractionation factor (17%); and (c) imposed changes of 5% in species-specific parameters



**FIGURE 4** Synthesis of the potential effects of various factors on the tuna  $\delta^{13}\text{C}$  temporal trend ( $\Delta\delta^{13}\text{C}$ ). Different combinations are possible (see text for more details) Tuna illustration Les Hata<sup>©</sup>SPC

indicating a shift in species composition (46%; Figure 4). While this is one potential scenario, in part informed and constrained by observations in the literature (Gregg et al., 2017; Rousseaux & Gregg, 2015), there is a multitude of permutations that may fit the observed trend. Nevertheless, changes in the carbon fixation fractionation factor ( $\epsilon_p$ ) had the largest effects in the simulations using both modeling approaches, and better accounted for the observed tuna trends than changes in productivity, the trophic fractionation factor ( $\epsilon_{tc}$ ), or even the known Suess effect.

## 4 | DISCUSSION

Our analysis revealed that changes in the biological component of the marine carbon cycle can be traced in the tissues of marine top predators. We observed substantial and widespread declines in tuna muscle  $\delta^{13}\text{C}$  values (by 0.8‰–2.5‰) in three tuna species across three ocean basins. Such a trend over a 16-year period has never been recorded in metabolic tissues of a marine predator. The use of two separate modeling approaches (sensitivity analysis and Bayesian inference) revealed that the parameter linked to phytoplankton carbon fractionation ( $\epsilon_p$ ) had the largest influence on the observed temporal trend in tuna muscle  $\delta^{13}\text{C}$  values. Our calculations then suggest that up to 60% of the decrease in tuna  $\delta^{13}\text{C}$  values seems to be due to a change in phytoplankton parameters in the Pacific Ocean, compared to only 27% in the western Indian Ocean. While our most robust dataset is from the Pacific Ocean, the same decreasing pattern in tuna  $\delta^{13}\text{C}$  values in all ocean basins (Pacific, Atlantic, and western Indian) suggests a widespread shift in marine plankton communities or a change in their physiology, but does not exclude other factors that may act in synergy (e.g., a change in productivity or in the trophic fractionation factor).

Previously reported temporal changes in  $\delta^{13}\text{C}$  values are generally attributed to the Suess effect or changes in marine productivity in various organisms and ecosystems (Fraile et al., 2016;

Newsome et al., 2007; Schell, 2001). For example, Schell (2001) found a significant long-term decline in  $\delta^{13}\text{C}$  values in inert baleen plates (~2.7‰) over a 30 year period (between 1965 and 1997) attributing this decline to a ~30%–40% decline in primary productivity in the Bering Sea. Cullen et al. (2001) proposed that part of the decrease observed by Schell (2001) was due to the Suess effect and due to the influence of changes in  $\text{CO}_2$  concentration on phytoplankton physiology (as shown in Figure 3b and described herein). In contrast, the Suess effect is relatively small over our time period (0.3‰; Figure 3a) and only explains ~12%–20% of the observed decrease in tuna muscle  $\delta^{13}\text{C}$  values (in the Atlantic and Pacific Oceans). Similarly, increasing  $\text{CO}_2$  concentrations only explain a small percentage (~2%–18%; Table 1) of the observed decrease in tuna muscle  $\delta^{13}\text{C}$  values, using the Bidigare et al. (1997) or Cassar et al. (2006) parametrizations for their effect on the carbon fixation factor fractionation.

After accounting for the observed Suess effect and changes in  $[\text{CO}_2]_{\text{aq}}$  availability, our model was used to explore how changes in phytoplankton growth rates and species composition can further reconstruct our observed declines in tuna muscle  $\delta^{13}\text{C}$  values. The sensitivity, Bayesian, and scenario analysis demonstrated that relatively small changes in phytoplankton community composition can lead to large declines in tuna  $\delta^{13}\text{C}$  values (Table 1; Analysis S1), while larger changes in productivity or the trophic fractionation factor would be needed. In our study, a 15% decrease in phytoplankton productivity cannot explain the decline we observed in tuna muscle  $\delta^{13}\text{C}$  values, even when combined with the Suess effect and the cumulative effect of increasing  $[\text{CO}_2]_{\text{aq}}$  on  $\epsilon_p$ . The change in productivity that we tested in this study is at the higher end of previously reported declines (typically ranging between 0% and 1.4% year<sup>-1</sup> at the regional to ocean basin scale; Behrenfeld, 2006; Gregg & Rousseaux, 2014; Joo, 2015). Other studies have shown no recent trends in global primary productivity (Gregg et al., 2017; Rousseaux & Gregg, 2017) and some regional increases have



been reported historically (e.g., Pacific Ocean; Karl, Bidigare, & Letelier, 2001). Even in model studies, projections of global marine NPP are highly uncertain with relative changes between -20% and +20% over the 21st century (Kwiatkowski et al., 2017).

The rate of decrease observed in our tuna  $\delta^{13}\text{C}$  values requires concomitant changes in phytoplankton species-specific parameters ( $\epsilon_f$  and  $b$ ). Young, Bruggeman, Rickaby, Erez, and Conte (2013) already reported evidence of change in the biological carbon isotopic fractionation by phytoplankton ( $\epsilon_p$ ) with significant increases between 1960 and 2010, in particular in the subtropics, where this change was the highest compared to other regions. The change in biological fractionation estimated through their model (i.e., a maximum of 0.4‰ in 16 years) is two to five times lower than our observations in tuna, depending on the ocean basin. Their study is based on a compilation of particulate organic carbon (POC) data from several transects mostly from the Atlantic Ocean with few data in our Pacific region, which could explain the differences between their model and our observations. However, they showed a time series of  $\delta^{13}\text{C}$  POC ( $\delta^{13}_{\text{POC}}$ ) values in the North Atlantic off Bermuda with a 2‰ decrease from 1980 to 2007 (i.e., ~1.2‰ decrease in 16 years). This result is similar to the 1.4‰ decrease in tuna values we observed in the Atlantic from 2000 to 2015.

Support for our hypothesis of a shift in marine plankton communities already exists (McMahon, McCarthy, Sherwood, Larsen, & Guilderson, 2015; Polovina, Howell, & Abecassis, 2008; Polovina & Woodworth, 2012; Rousseaux & Gregg, 2015). Diatoms are predicted to decrease in abundance in response to increased seawater stratification with a reported decline of 1.22% year<sup>-1</sup> in the North Pacific (Rousseaux & Gregg, 2015). Such a reduction in the abundance of diatoms, a <sup>13</sup>C-rich carbon source in marine food webs (Fry & Wainright, 1991), is expected to decrease the  $\delta^{13}\text{C}$  values of consumers, and this diatom contribution has been emphasized in a recent model of phytoplankton  $\delta^{13}\text{C}$  variations in the global ocean (Magozzi et al., 2017). Tuerena et al. (2019) also recently found that cell size was the primary determinant of  $\delta^{13}_{\text{POC}}$  in the South Atlantic subtropical convergence zone and predicted that carbon isotopic fractionation will increase in the future, leading to lower  $\delta^{13}_{\text{POC}}$  that may propagate through the food web. A decrease in the abundance of coccolithophores, another <sup>13</sup>C-rich carbon source, might also explain some of the tuna  $\delta^{13}\text{C}$  trend. However, Rivero-Calle, Gnanadesikan, Castillo, Balch, and Guikema (2015) found an increase in the occurrence of coccolithophores in the North Atlantic, but data are not available at a global scale. Ocean basin differences found in our study in the temporal slopes in tuna  $\delta^{13}\text{C}$  values between the Indian Ocean (0.8‰) and the Atlantic and Pacific Oceans (from 1.4‰ to 2.5‰) could be due to a combination of several factors. The magnitude of the Suess effect may vary regionally (Quay et al., 2016). Changes in phytoplankton communities or physiology are also dependent on regional-scale processes (Gregg et al., 2017; Siegel et al., 2013). The use of spatially resolved models of the ocean <sup>13</sup>C cycle would help to understand the regional differences (Tagliabue & Bopp, 2008).

## 5 | CAVEATS AND LIMITATIONS

Phytoplankton have many strategies to take up carbon as a function of growth conditions that could affect fractionation. For example, a decrease in bicarbonate uptake or carbon-concentrating mechanism activity in general would be predicted to increase the apparent fractionation. Our predictions should therefore be interpreted with caution as isotopic fractionation is not a single function of  $\mu/\text{CO}_2$ , even within a single phytoplankton species (Cassar & Laws, 2007; Cassar et al., 2006). Furthermore, other permutations of  $\epsilon_f$  or  $b$  may fit the observed decrease. However, this parametrization together with the Bayesian inference demonstrates how small changes in phytoplankton community composition or physiology may influence tuna muscle  $\delta^{13}\text{C}$  values.

We also note that regional variations cannot be captured by the time series of  $\delta^{13}\text{C}_{\text{aq}}$  at Station ALOHA (Hawaii, Pacific). Long-term declines of  $\delta^{13}\text{C}_{\text{aq}}$  values due to the combination of the Suess effect, vertical mixing, and primary production (residual carbon pool after POC production) have been documented at other monitoring stations, with varying effects according to region and latitude, in particular in the southern regions (King & Howard, 2004). However, both instrumental and proxy records of  $\delta^{13}\text{C}_{\text{aq}}$  indicate a consistent average decrease per year of 0.027‰ at five Pacific stations from Hawaii to American Samoa since 1980 (corresponding to an approximately 0.4‰ decline in our 16 year period; Wu, 2018). Furthermore, Gruber et al. (1999) compared the  $\delta^{13}\text{C}_{\text{aq}}$  trends in several oceanic regions and found that the highest decrease of 0.025‰ was in the subtropical gyres (Bermuda and Hawaii) and the lowest in the equatorial upwelling region of the Pacific (0.015‰), with the Indian Ocean displaying a decrease of 0.020‰ per year. Therefore, the predicted ranges in all oceanic regions of 0.2‰–0.4‰ decrease over our study period of 16 years are too small to explain the 1.8‰ average decline in tuna  $\delta^{13}\text{C}$  values.

Other factors, related to food web or dietary processes, could also influence the tuna  $\delta^{13}\text{C}$  trend. Size differences in sampled tuna through time could introduce a bias, but no consistent relationship between tuna size and  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values or any size changes with time among tuna species were observed (Figures S5 and S6). While decadal shifts in the diet of yellowfin tuna have been recorded in the eastern Pacific Ocean from the 1990s to the 2000s (Olson et al., 2014), the similar  $\delta^{13}\text{C}$  slopes observed for the three tuna species in our study seem inconsistent with changes in foraging location or diet. A shift in the tuna foraging range or timing could also be argued to explain the observed decrease in tuna  $\delta^{13}\text{C}$  values as there are large spatial and intraannual variabilities in the  $\delta^{13}\text{C}$  values of phytoplankton (Magozzi et al., 2017). Similarly, if all tuna foraged deeper on more mesopelagic prey that had lower  $\delta^{13}\text{C}$  values than surface prey ( $\delta^{13}\text{C}_{\text{DIC}}$  is known to decrease with depth; Quay et al., 2003), tuna  $\delta^{13}\text{C}$  values would decrease. However, the slope of this decrease would vary among species given that yellowfin tuna mainly inhabit surface waters while bigeye tuna mostly forage in mesopelagic waters (Olson et al., 2016). A decrease in the trophic

fractionation factor  $\epsilon_{fc}$  would reduce tuna  $\delta^{13}\text{C}$  values through time (Table 1; Figure 4). Various processes including a change in food chain length, food web structure, quality of food, or tuna metabolism (Barnes, Sweeting, Jennings, Barry, & Polunin, 2007) could alter  $\epsilon_{fc}$ . A change in the overall trophic fractionation factor could therefore occur at multiple levels of the food web, driven or not by changes at the base of the food web. To our knowledge, there are no data available in the literature to explore this further. However, while we cannot rule out the possibility that changes in food web structure are negated by changes in source  $^{15}\text{N}$  (e.g., denitrification vs.  $\text{N}_2$  fixation; Deutsch et al., 2014; Somes et al., 2010), we did not see temporal changes in tuna  $\delta^{15}\text{N}$  values at global or ocean basin scales, suggesting little change in food chain length or structure.

Finally, our dataset has some limitations inherent to the sampling and we acknowledge that our most robust analysis is for the Pacific Ocean that covers a large area with many individuals by year and species (Table S1). More data over broad reaches of the Atlantic and Indian Oceans are needed to provide robust estimates of biological changes in these oceans.

## 6 | CONCLUDING REMARKS

We showed that  $\delta^{13}\text{C}$  values of metabolically active tissues of mobile marine predators likely reflect recent changes at the base of marine food webs. We detected a substantial worldwide decrease in tuna  $\delta^{13}\text{C}$  values over the 2000–2015 period which can be related to various processes known to influence ocean carbon cycling in the global oceans. Our analysis suggests that phytoplankton species (e.g., diatoms) that undergo a larger fractionation of carbon during photosynthesis (and thus have higher  $\delta^{13}\text{C}$  values) have been decreasing over recent decades or that these phytoplankton communities altered their physiologies. While we cannot rule out a widespread decline in phytoplankton productivity, we showed that even a large (>15%) decline would have a small impact on tuna  $\delta^{13}\text{C}$  values and cannot fully explain the observed global trend. While recognizing that a concomitant shift at higher levels of food webs (change in the trophic fractionation factor or in tuna diet or physiology) could occur and that more tuna carbon isotope data are needed from the Atlantic and Indian Oceans, the present study expands our understanding of the main factors that affect the isotopic values of top predators and provides a framework to interpret and model carbon cycling at regional to global scales. New observational or modeled data that provide estimates of periodic changes in marine plankton communities will enable our model to provide estimates of the other contributing factors. Finally, the framework presented here, through the study of tuna carbon and nitrogen isotopes values, could support development of a useful essential ocean variable (EOV) for implementation within a global ocean observing system to document complex ecosystem changes at regional to global scales and over relatively short timescales (decades to centuries). The use of predator isotopes as an EOV would complement regional efforts to acquire in situ measurements of plankton abundance and diversity (Miloslavich et al., 2018).

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## CONFLICT OF INTEREST

The authors declare no competing interests.

## AUTHOR CONTRIBUTIONS

A.L., A.R., N.C., and H.P. analyzed the data and interpreted the results with the help of B.F. A.R. and D.E.P. performed the statistical analysis. A.L., N.C., and H.P. wrote the manuscript with the help of A.J.H. and A.R. All authors contributed to and provided feedback on various drafts of the paper.

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## SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

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